# Harvest, density and reproductive characteristics of North American walleye populations 

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#### Abstract

A synopsis of walleye population characteristics from North American lakes is presented. Harvest, density and reproductive data from the primary literature, agency reports and personal communications were summarized, and include: sport fishing harvests, exploitation rates, population densities, relative fecundity, and age to $50 \%$ maturity. Empirical relationships between yield, population size, lake area, relative fecundity, age to maturity and growing degree-days are described. Factors affecting these empirical relationships are also discussed. Quartiles were used to describe the frequency distributions of harvest, yield and density parameters. Managers can use these empirical relationships and descriptive statistics as comparative diagnostic tools for interpreting the status of their walleye fishery. We feel this is a useful approach because managers are often having to make decisions regarding their fishery with minimal information. The interpretive value of minimal data can be enhanced when comparative information is available.


## 1. Introduction

For many fisheries, especially those with limited access, such as in Northern Ontario (above $50^{\circ}$ latitude), there is limited information with which to make management decisions. Reduced funding for fishery management programs, prevents further reduction of uncertainty in the decision making process. However, a limited amount of harvest information along with some data describing the physico-chemical features of a water body is frequently available
to assess a fishery.
The interpretive value of limited data can be greatly enhanced when comparative information is available. The purposes of our paper are: (1) to synthesize available harvest, density, exploitation and reproduction data from numerous walleye populations over a large geographical area; (2) describe empirical relationships between these parameters; and (3) discuss factors which determine stock status. Although uncertainty is implicit in this approach, due to unexplained variances inherent within and


Fig. 1. Frequency distribution of walleye angling yields (kg/ha) from 168 North American waters.
between aquatic communities, this review provides an improved understanding of the limits that can be expected for various population characteristics, and how they relate to each other.

## 2. Methods

Walleye harvest and density data was compiled from the primary literature, internal agency reports and extensive personal communication with fisheries workers from the Ontario Ministry of Natural Resources and other resource agencies. We analysed data from a range of walleye populations to observe their variation, common properties and dynamics in various habitats. Quartiles were used as descriptive statistics because val-

Table 1. Harvest characteristics of North American walleye populations showing the 25,50 , and 75 percent quartiles for angling exploitation rates (percent), adult density (number/ha), harvest weight (kg) and yield (kg/ha). Also shown are sample sizes ( $n$ ) and range of values.

|  | Quartiles |  |  |  |  |  |
| :--- | :---: | :---: | :---: | ---: | :---: | :---: |
|  | 25 | 50 | 75 | $n$ | Range |  |
| Exploitation Rates | 14 | 21 | 25 | 46 | $3-55.6$ |  |
| Adult Density | 7.8 | 14.8 | 23.9 | 85 | $0.1-168$ |  |
| Harvest Weight | 0.48 | 0.58 | 0.67 | 113 | $0.26-1.18$ |  |
| Yield | 0.50 | 1.24 | 2.95 | 168 | $0.01-49.6$ |  |



Fig. 2. Plot of annual walleye angling harvest (kg) against lake area (ha) for 92 single-year, and 75 multiyear (two or more years) observations.
ues of yield, population density and exploitation rates have skewed distributions that do not conform to standard probability models. Thus, the frequency distribution is divided into equal quartiles, 25,50 and $75 \%$. The $50 \%$ quartile is the median.

The number of observations varied among populations, and estimates of total harvest included both commercial and angling yields when available. About 95\% of the walleye populations were harvested by angling only. Estimates of population and yield characteristics varies between water bodies. We assumed that both single-year and multi-year observations reflected long-term averages and approximates sustainable yields at least as they are presently impacted by our cultural practices. The degree of error associated with this assumption will only become apparent with time, hopefully it will be minimal and manageable.

## 3. Results

### 3.1. Yield

The distributions of walleye yields is highly skewed to the right (Fig. 1). The mode of the distribution is for yields less than $1 \mathrm{~kg} / \mathrm{ha}$ ( $42 \%$ of the sample). The $25 \%$ quartile corresponds to a yield of $0.50 \mathrm{~kg} / \mathrm{ha}$, the $50 \%$ quartile (median) is $1.24 \mathrm{~kg} / \mathrm{ha}$, and the $75 \%$ quartile is $2.95 \mathrm{~kg} / \mathrm{ha}$ (Table 1 ).


Fig. 3. Frequency distribution of adult walleye population densities (numbers/ha) from 85 North American waters.

There is a significant $(p<0.01)$ relationship between total walleye yield ( kg ) and lake area (ha). We regressed 168 observations of log yield against log area (Fig. 2) and obtained the following significant $\left(p<0.01, R^{2}=0.59\right)$ relationship:

$$
\begin{equation*}
Y I E L D=1.81 A R E A^{0.931} \tag{1}
\end{equation*}
$$

Out of 167 observations, 92 (55\%) consisted of single-year yield estimates, and 75 (45\%) had two or more years of data. We fitted separate regression lines to single- and multi-year data, and found that the slopes were not significantly different (ANOVA $F$-test, $p<0.01$ ). Therefore, we combined the data to generate equation 1 .

### 3.2. Population size

The frequency distribution of population density expressed as number of walleye per hectare, is also skewed to the right (Fig. 3). The mode of the distribution occurs at a density of less than 10 walleye/ha ( $34 \%$ of the sample). The 25,50 and $75 \%$ quartiles for walleye density are $8.1,14.8$ and 23.9 walleye/ha respectively (Table 1).


Fig. 4. Plot of adult walleye population size against lake area (ha) for 81 North American waters.

The size of a walleye population increases with lake area (Fig. 4). A log-log regression of lake area (ha) with population size of adult walleye $(P O P)$ for a sample of 81 lakes resulted in the following significant ( $p<0.01, R^{2}=0.75$ ) relationship:

$$
\begin{equation*}
P O P=27.31 A R E A^{0.79} \tag{2}
\end{equation*}
$$

### 3.3. Reproduction

Fecundity is defined as the number of eggs present in a fish prior to spawning. It can be expressed as the total number of eggs per female (absolute fecundity), or number of eggs per kilogram of body weight (relative fecundity). Table 2 shows relative fecundities of walleye populations, and the number of Growing Degree-Days (GDD) for those locations. GDD are defined as the number of degree-days above $5 \mathrm{Cel}-$ sius and is a standard measurement reported by Environment Canada's Weather Service. GDD have been used to quantify energy available for walleye growth and reproduction (Colby \& Nepszy 1981).

The mean relative fecundity $(R F)$, expressed in eggs per kilogram, was plotted against $G D D$ (Fig. 5). The resulting relationship is a statistically-significant ( $p<0.01, R^{2}=0.67$ ) straight line:

Table 2. Fecundity of walleyes, expressed as number of eggs per kilogram, and growing degree-days (GDD), from North American waters.

| Waterbody | Location | Year of study | $G D D$ | Range | Mean | Ref. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| George Lake | ON | 1983 | 1200 |  | 39267 | 1 |
| Lower Tweed Lake | ON | 1983 | 1200 |  | 39029 | 1 |
| Upper Kesagami Lake | ON | 1983 | 1200 |  | 38540 | 1 |
| Wakwaycwkastic Lake | ON | 1983 | 1200 |  | 46130 | 1 |
| Kowashkagama River | ON | 1982 | 1250 | 33 102-59 064 | 43922 | 2 |
| Onaman Lake | ON | 1982 | 1250 | 35914-65035 | 51792 | 3 |
| Henderson Lake | ON | 1980 | 1287 |  | 41300 | 4 |
| Henderson Lake | ON | 1981 | 1287 |  | 52500 | 4 |
| Henderson Lake | ON | 1982 | 1287 |  | 49800 | 4 |
| Henderson Lake | ON | 1983 | 1287 | 29 822-83 286 | 48900 | 4 |
| Henderson Lake | ON | 1984 | 1287 | 44 750-56216 | 50900 | 4 |
| Henderson Lake | ON | 1985 | 1287 | 45 068-63783 | 51500 | 4 |
| Lac Des Mille Lacs | ON | 1982 | 1287 | 34 700-46860 | 58360 | 5 |
| Savanne Lake | ON | 1980 | 1287 | 28966-48666 | 39690 | 4 |
| Savanne Lake | ON | 1981 | 1287 | 32 850-50 579 | 39800 | 4 |
| Savanne Lake | ON | 1982 | 1287 | 22 681-54 016 | 42400 | 4 |
| Savanne Lake | ON | 1983 | 1287 | $36145-56411$ | 43200 | 4 |
| Savanne Lake | ON | 1984 | 1287 | 28138-69 698 | 49300 | 4 |
| Savanne Lake | ON | 1985 | 1287 | 33 510-60 948 | 43580 | 4 |
| Home Lake | MB | 1977 | 1400 | 34 260-51590 | 40513 | 6 |
| Wapun Lake | MB | 1977 | 1400 | 24 674-61 095 | 52508 | 6 |
| Lake Winnipeg (north end) | MB | 1975 | 1419 | 27 354-106 379 | 59233 | 7 |
| Lake Winnipegosis | MB | 1987 | 1440 |  | 50487 | 7 |
| Lake of the Woods | MN | 1941 | 1657 |  | 50000 | 8 |
| Little Cutfoot Sioux Lake | MN | 1954 | 1662 | 48840-73 700 | 65239 | 9 |
| Lake Winnipeg (south end) | MB | 1981 | 1688 | 26 394-69 814 | 50665 | 7 |
| Lake Nipissing | ON | 1984 | 1780 | 25 275-83 414 | 47387 | 10 |
| Balsam Lake | ON | 1987 | 1800 | 43 600-63690 | 53094 | 11 |
| Pigeon Lake | ON | 1987 | 1800 | $55926-87604$ | 68051 | 11 |
| Lake Gogebic | MI | 1947 | 1864 | 57 922-67 797 |  | 12 |
| Muskegon River | MI | 1947 | 1864 | $65778-95955$ |  | 12 |
| Big Sand Lake | MN | 1900 |  | 61370 |  | 13 |
| Moon River | ON | 1900 |  | 43 925-100 313 | 65083 | 14 |
| Escanaba Lake | WI | 1979-81 | 1901 | $48441-74161$ | 67914 | 15 |
| Otter Tail Lake | MN | 1910 |  | 45298 |  | 16 |
| Bay of Quinte (L. Ontario) | ON | 1962 | 2000 | 57096-100 031 | 73486 | 17 |
| Fox River | WI | 1986 | 2144 | 31000-92600 | 51600 | 18 |
| Lake Winnebago | WI | 1964-67 | 2187 | 63 441-96 116 |  | 19 |
| Lake Erie (W. basin) | OH | 966 | 2600 | 41 191-96914 | 61149 | 20 |
| Lake Erie (W. basin) | OH | 990-91 | 2600 | 52 980-147160 | 84710 | 21 |
| Lake St. Clair | ON | 1977 | 2500 |  | 87397 | 22 |
| Lake Erie (E. basin) | OH | 1966 | 2300 | 56 314-123 249 | 82700 | 20 |
| Columbia River (John Day Pool) | OR | 1981 | 2730 | 69000-101000 | 82900 | 23 |
| Lake Meredith | TX | 1968-71 | 3691 | 36500-72 200 | 52000 | 24 |
| Center Hill Reservoir | TN | 1965-66 | 4078 | 37954-143827 | 64715 | 25 |
| Norris Reservoir | TN | 1939-40 | 4078 | 28 415-32 727 | 29700 | 26 |
| Mississippi River |  |  |  | 50600-110100 |  | 27 |
| Utah Lake | UT |  |  | 27 900-52 562 | 47410 | 28 |
| Wisconsin waters | WI |  |  | 28600-99 000 |  | 29 |

1. Armstrong and Jolkowski (unpub. data), 2. Sobchuk (unpub. data), 3. Walroth (unpub. data), 4. Baccante and Reid (1988), 5. Fruetel (unpub. data), 6. Babuluck (pers. comm.), 7. Lysak (pers. comm.), 8. Carlander (1945), 9. Johnson (1971), 10. Jorgensen (unpub. data), 11. Deacon (unpub. data), 12. Eschmeyer (1950), 13. P. Jacobson (pers. comm.), 14. Winterton (1975), 15. Serns (1982), 16. D. Schreiner (pers. comm.), 17. Payne (1963), 18. Balcer et al. (1986), 19. Priegel (1970), 20. Wolfert (1969), 21. Muth and Ickes (1993), 22. MacLennan (pers. comm.), 23. Maule and Horton (unpub. data), 24. Kraai and Prentice (1974), 25. Muench (1966), 26. Smith (1941), 27. Nord (1967), 28. Arnold (1960), 29. Niemuth et al. (1966).


Fig. 5. Plot of relative fecundity (eggs/kg) against growing degree days (above $5^{\circ} \mathrm{C}$ ) for 45 North American walleye populations.

$$
\begin{equation*}
R F=24.668 G D D+14514 \tag{3}
\end{equation*}
$$

The three outliers in Fig. 5 are excluded from the regression. These populations are from the extreme southern range (Tennessee and Texas) of the species and are subject to physiological constraints on reproduction, mentioned in the discussion section of this paper.

Beverton (1987) tabulated data from Colby and Nepszy (1981) and plotted age to $50 \%$ maturity $\left(T_{\mathrm{m}}\right)$ against $G D D$. We added additional data and fitted a line to quantify the relationship (Fig. 6). The resulting best-fit is a power curve of the form:

$$
\begin{equation*}
T_{\mathrm{m}}=3184.72 G D D^{-0.871} \tag{4}
\end{equation*}
$$

The regression is statistically significant ( $p<0.01, R^{2}=0.87$ ).

### 3.4. Exploitation rates

Angling exploitation rates for walleye from various sources were compiled and tabulated (Table 3). The median exploitation rate in a sample of 46 observations is $21 \%$, and range from 3 to $55.6 \%$ (Table 1). The frequency distribution of the exploitation rates show a skewed distribution with a mode


Fig. 6. Plot of age to 50 percent maturity against growing degree days for 23 North American walleye populations.
at the 20-30\% category (Fig. 7).
We plotted exploitation rates against Growing Degree-Days (GDD), for 16 lakes (Fig. 8). On the graph we have used different symbols to indicate which fisheries have "collapsed" or severely depleted, and which have been able to support their respective level of exploitation (sustainable). The data suggest that there may be a curvolinear relationship between exploitation and energy, with an upper threshold beyond which the population will not sustain itself. The line on the graph is not fitted statistically, rather, it represents our best guess, based on long-term data from a few lakes, of maximum, long-term sustainable exploitation rates.

## 4. Discussion

### 4.1. Yield

Our results indicate that lake area provides adequate first-order estimation of walleye yield. Yield in kilograms and lake area in hectares both increase proportionally over a logarithmic scale. Thus, larger lakes produce larger walleye yields. However, we must caution that, yield estimates may not reflect

Table 3. Walleye exploitation rates (percent) as reported in various angling fisheries.

| Waterbody | Location | GDD | Area (ha) | Mean | Range | Period | Duration (seasons) (*months) | Ref. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Abamasagi Lake | ON |  | 1526 | 32 |  | 1981 | 1 | 1 |
| Big Sand Lake | MN | 1900 | 664 | $21.9{ }^{\text {a) }}$ |  | 1990-92 |  | 2 |
| Big Sand Lake | MN | 1900 | 664 | $22.3{ }^{\text {b) }}$ |  | 1990-92 |  | 2 |
| Branched Oak Lake | NE |  | 728 | 22.1 | 19.9-24.2 | 1979-80 | 2 | 3 |
| Cass, Andrusia, Big Wolf Lakes | MN |  | 7369 | $15.8{ }^{\text {a) }}$ |  | 1971-75 |  | 4 |
| Cass, Andrusia, Big Wolf Lakes | MN |  | 7369 | $26.0^{\text {b) }}$ |  | 1971-75 |  | 4 |
| Cutfoot Siouz Lake | MN | 1662 | 953 | 17 | 11-22 | 1957-58 | 2 | 5 |
| DuBay Lake | WI |  | 2692 | 18.8 |  | 1983-84 | 1 | 6 |
| Escanaba Lake | WI | 1901 | 119 | 25 |  | 1953-82 | 30 | 7 |
| Fife Lake | MI |  | 251 | 5.1 |  | 1964-65 | 2 | 8 |
| Flambeau Flowage | WI |  | 5792 | 21 | 2-43 | 1975-76 | 2 | 9 |
| Gogebic Lake | MI | 1864 | 5986 |  | 2-6 | 1947, 1976 |  | 10 |
| Goulais River | ON |  |  | 29 |  | 1962, 1964 | 15* | 11 |
| Green Bay (L. Michigan) | Ml |  |  | 4.1 | 1.6-6.3 | 1957-63 |  | 12 |
| Henderson Lake | ON | 1250 | 151 | 55.6 | 49-66.2 | 1980-82 | 3 | 13 |
| Home Lake | MB | 1400 | 169.8 | 22.5 | 16-29 | 1979-80 | 2 | 14 |
| Hoover Reservoir | OH |  | 1143 | 29 |  | 1967 | 1 | 15 |
| Inland Waterway (Burt Lake) | MI |  | 6758 |  | 7-18 | 1950's-75 |  | 10 |
| Jewett Lake | MI |  | 5.2 | 8.9 |  | 1979-82 | 4 | 12 |
| Kabetogama Lake | MN | 1241 | 10428 | 23 |  | 1984-85 | 2 | 16 |
| Kenogamissi Lake | ON |  | 2499 | 10.5 |  | 1974 | 1 | 17 |
| Lake Erie | ON | 2600 | 2569000 | 18.8 | 14.9-23.9 | 1977-86 |  | 18 |
| Lake of the Woods | MN | 1657 | 128294 | 13.7 |  | 1982 |  | 19 |
| Many Point Lake | MN |  | 694 | 27 | 21-33 | 1955-57 | 3 | 20 |
| Meredith Lake | TX | 1905 | 4452 | 5 | 1.6-9.6 | 1986-93 | 8 | 21 |
| Michigan Average | Ml | 1900 |  | 21 |  |  |  | 22 |
| Mille Lacs | MN | 1900 | 53419 | 25 |  |  |  | 23 |
| Mississauga River | ON |  |  | 14 |  | 1965 | 1 | 24 |
| Moon River | ON | 1900 |  | $15.6{ }^{\text {c }}$ | 11-22 | 1969-72 | 4 | 25 |
| Muskegon River | Ml | 1864 |  | 16.1 | 9.7-25.5 | 1947-50 | 3 | 12 |
| Nipigon Bay, (L. Superior) | ON | 1200 |  | 18 | 7-34 | 1955-57 | 3 | 26 |
| Okoboji Lake (East and West) | IA | 2250 | 2301 | 19.4 |  | 1991 | 1 | 27 |
| Oneida Lake | NY | 2500 | 20640 | 23 | 10-47 | 1957-59 | 3 | 28 |
| Otter Tail Lake | MN | 1900 | 5554 | 34 |  | 1983 |  | 29 |
| Pike Lake | WI |  | 211 | 21.6 | 20.7-22.4 | 1959-62 | 3 | 30 |
| Sallie Lake | MN |  | 504 | 9.3 |  | 1954 |  | 31 |
| Savanne Lake | ON | 1250 | 364 | 16.6 | 11.9-23.6 | 1983-86 | 4 | 12 |
| Spirit Lake | IA | 2250 | 2288 | 22 | 15-29 | 1947, 1954, 1991 | 13 | 32 |
| St. Louis River Estuary | MN |  |  | 8.1 |  | 1980-82 |  | 33 |
| Touchwood Lake | AB | 1100 | 2900 | 3 | 2-4 | 1991 | 1 | 34 |
| Upper Chukuni River (Red Lake) | ON | 1348 |  | 25.9 | 22.5-29.2 | 1984-85 | 2 | 35 |
| Vermillion Lake | MN |  | 16413 | 5 |  | 1940 |  | 36 |
| Wapun Lake | MB | 1400 | 212.6 | 52.5 | 52-53 | 1979-80 | 2 | 14 |
| Whitefish Lake | ON | 1200 | 3015 | 12 | 10-30 | 1989-90 | 2 | 37 |
| Winnibigoshish Lake | MN |  | 23693 | 10.2 |  | 1937-39 |  | 38 |
| Winnibigoshish Lake | MN |  | 23693 | $26.1^{\text {a) }}$ |  | 1975-77 |  | 39 |
| Winnibigoshish Lake | MN |  | 23693 | $23.1{ }^{\text {b) }}$ |  | 1975-77 |  | 39 |
| Wolf Lake | WI |  | 159 | 8.6 | 5.7-12.3 | 1972-76 | 4 | 40 |
| Wolf Lake | AB | 1100 | 3150 | $22.0{ }^{\text {b) }}$ | 21-23 | 1991 | 1 | 35 |

${ }^{\text {a) }}$ Mature females, ${ }^{\text {b) }}$ Mature males, ${ }^{\text {c) }}$ An additional 12 percent due to commercial fishing, ${ }^{\text {d) }}$ An additional 3 percent due to commercial fishing.

1. Sobchuk (1981), 2. Jacobson (pers. commun.), 3. Schainost, S. (1983), 4. Strand (1980), 5. Johnson and Johnson (1971), 6. Hauber, A. (1989), 7. Serns (1985), 8. Schneider (1969), 9. Lealos and Bever (1982), 10. In: J. C. Schneider (1978), 11. Rose (1984), 12. Schneider (1977), 13. Reid and Momot (1985), 14. Babaluck, J. (F\&O pers. comm.), 15. Erickson and Stevenson (1967), 16. Kallemeyn, L. W. (1986), 17. Deyne (1983), 18. OMNR unpub. report, 19. Payer et. al. (1987), 20. Olson (1958), 21. Munger, C. (TX P.F. \& W., pers. comm.), 22. Schneider (1978), 23. Schupp (MNDNR pers. comm.), 24. Payne (1965), 25. Winterton (1975), 26. Ryder (1968), 27. Larscheid, J. (1992), 28. Forney (1967), 29. Schreiner (1987), 30. Mraz (1968), 31. Olson (1955), 32. Rose (1955); Larscheid, J. (1992), 33. Osborn et. al. (1991), 34. Sullivan, M. (AB Nat. Res., per. comm.), 35. Weilandt (1984), 36. Carlander (1941), 37. Fruetel (1994), 38. Stoudt and Eddy (1939), 39. Osborn et. al. (1985), 40. Serns (1981).


Fig. 7. Frequency distribution of angling exploitation rates (percent) for 46 North American walleye populations.
populations in equilibrium. In reality, some of these lakes may not be in equilibrium and present yields may exceed long-term stable or sustainable yields. This reiterates the importance of long-term monitoring studies in fisheries science.

The strong relationship between walleye yield and lake area is consistent with other studies which correlate fish productivity with lake morphometry (Leach et al. 1987). However, in the literature review provided by Leach et al. (1987), it is evident that, although walleye yield has been correlated to macrobenthos production (Matuszek 1978), climate and effort (Schlesinger \& McCombie 1983), none of the studies deal specifically with walleye yield and lake area relationships.

Lake area has been correlated with the number of fish species on a global (Barbour \& Brown 1974), and regional scale (Harvey 1978). Rounsefell (1946) found an inverse relationship between fish yield and lake area, that is, large lakes produce less fish per unit area than small lakes. However, because large lakes are generally deeper than small lakes (Hayes 1957), yield is really a function of depth as well as area. Also, smaller, shallower lakes have a larger proportion of the substrate in the euphotic zone, which is important for fish production. Conversely, most of


Fig. 8. Plot of angling exploitation rates that have been sustained or have resulted in the collapse of walleye from a sample of 27 North American waters. The line was not fitted statistically, it represents our estimate of maximum sustainable rates over the energy cline.
the substrate in the larger, deeper lakes may be in the profundal zone, which is aphotic, and acts as a sink for nutrients, making them unavailable to production. Rawson (1952) hypothesized that increasing lake depth constrains fish production beyond the limit where increase in area has any effect.

Rousenfell's (1946) observation of the inverse relationship between fish yield per unit area and lake area, has been confirmed by other researchers (Jenkins \& Morais 1971, Carlander 1977). Although it has been recognized for years that small lakes produce more yield per unit area than large lakes, this cornerstone of fisheries management has spurred a healthy dose of controversy. Youngs and Heimbuch (1982) suggested that there is a spurious correlation between lake area and yield expressed in kilograms per hectare, since the variable area occurs in both the dependent and independent variable. More recently, Jackson et al. (1990), criticized the use of ratios, such as the $M E I$, in aquatic research applications because the distribution of the ratio variables is not known which, they feel, leads to errors in prediction.

Rempel and Colby (1991) developed the morp-
hoedaphic model (MEM) in response to criticism of the MEI (Jackson et al. 1990), and use the MEM to predict fish yield in lakes. The MEM consists of two morphometric (area and volume) and one edaphic variable (TDS). Basin morphometry alone accounts for about $97 \%$ of the variation in fish harvest. Although lake volume and TDS only account for an additional 2.5 and $0.64 \%$ of the variability in harvest, the authors included them to explain variations in lakes with atypical basin morphometry and $T D S$ levels.

Despite criticism, the $M E I$ has proved to be a useful tool for predicting fish yields in lakes. It generally accounts for about 60 to $78 \%$ of the variation in fish yields. Youngs and Heimbuch (1982) showed that, the $M E I$ is a valid predictor for two reasons. First, large lakes produce greater fish yields than small ones. Second, large lakes are generally deeper, thus mean depth in the $M E I$ acts as a surrogate for surface area. These authors show that area alone can be a more powerful predictor of yield. They found that area accounted for about $94 \%$ of the variation in fish yield in three sets of data reported in the literature. They also found that the variation accounted for by the regression increased to only $95 \%$ when $T D S$ was added. Surface area accounts for $61 \%$ of the variation in walleye yield in the set of data we reported here. The variation about the yield-area regression line can be attributed to a number of factors, such as, community structure, habitat type, nutrient loading, methods of harvest, and fishing effort characteristics.

The first three factors were dealt with by Rempel and Colby (1991). They provide good examples of how each affects yield. Elsey and Thompson (1977) provide evidence of how different methods of harvest can influence walleye yields. Schlesinger and McCombie (1983) recognized that failure to partition effort by species in a multi-species fishery can add variance to yield predictions. Future research should be directed towards further understanding of how each of the above-mentioned factors affect walleye yield.

### 4.2. Population size

The relationship between lake area and walleye population size is analogous to the yield-area relationship, that is, larger lakes have more walleye but a less dense population when divided by area or
volume. Hansen (1989) reported a relationship between walleye abundance (numbers) and lake area (acres), and found that the latter explained $65 \%$ of the variation in walleye abundance in a set of 104 Wisconsin lakes. We did not have abundance data on all of Hansen's (1989) lakes, but used 19 from lakes in Hansen et al. (1991) in our set of 81 lakes and found that area explained $76 \%$ of the variation in walleye abundance. The abundance figures we used were all obtained from mark-recapture estimates, mostly reported in unpublished literature, obtained through personal communications with other researchers.

An empirical relationship between walleye abundance and lake area has proven to be helpful in estimating walleye abundance in northern Wisconsin lakes when other information was lacking (Hansen 1989). This approach provides inexpensive estimates of walleye abundance based on lake area. Safety factors can then be derived to allow setting of harvest quotas based on these estimates of walleye abundance (Hansen et al. 1991).

Between-lake differences in adult walleye density can be attributed to various parameters, such as, energy input, community type, habitat suitability, etc. These factors directly or indirectly affect walleye survival from the egg to the adult stages. Table 4 shows an example of how walleye density differs in two lakes, Savanne (Ontario) and Escanaba (Wisconsin). Savanne, a relatively unexploited lake, has an area of 364 ha , and is typical of many smallto medium-size mestrophic walleye lakes in Northern Ontario. Escanaba has an area of 119 ha , and has supported heavy angling pressure since the 1940s (Kempinger \& Carline 1977). Table 4 shows that Escanaba Lake has about three times more egg production per hectare than Savanne Lake. The higher fecundity is likely a result of greater energy input, 1367 $G D D$ in Savanne and $1901 G D D$ in Escanaba (Table 5). The survival from egg to each age group is similar in both lakes, but since Escanaba walleye have much higherrelative fecundity ( 67914 eggs $/ \mathrm{kg}$ ) than those in Savanne (39690 eggs $/ \mathrm{kg}$ ) (Table 2), the resulting adult density is higher.

Availability of ideal habitat is also a very crucial factor which determines population density. The importance of habitat as an indicator of successful establishment and reproduction of walleye populations has been well documented in the literature. A recent summary of these observations
is provided by Colby et al. (1991). Equally important is the accuracy of the estimate of adult abundance. Hansen et al. (1991) provide a detailed summary of factors which can affect this accuracy, such as changes in the vulnerability of the fish during the sampling period (see also Colby \& Baccante 1996). Hansen et al. (1991) also caution that if the estimate is used for management purposes, for example to set harvest quotas, the accuracy of population estimates declines over time from the year the estimate was obtained to the year it was used. They used data from Escanaba Lake, which has abundance estimates from 1953 to the present, to test the accuracy of population estimates over time. Hansen et al. (1991) stress the importance of long-term series of abundance estimates as a calibration tool. This has allowed them to corroborate their results from other lakes where abundance estimates were based on irregular sampling.

### 4.3. Reproduction

The shape of the curve in Fig. 6 suggests that relative fecundity $(R F)$, expressed as the mean number of eggs per kilogram is higher in systems with greater insolation. However, $R F$ decreases when
$G D D$ values reach values greater than 3500 . This limiting effect is indicative of physiological constraints to reproduction which have been welldocumented by Colby and Nepszy (1981). They have shown that, ovary maturation may not occur in environments receiving too much energy. They cite data by Hokanson (1977) which indicate that a minimum winter temperature of $10^{\circ} \mathrm{C}$ is near the upper limit for maturation of gonads in yellow perch and walleye. The southern-most walleye population in Ontario is in western Lake Erie, with $2300 G D D$, thus reproduction is not limited by lack of cool temperatures during the maturation cycle.

Describing how $R F$ changes over an energy cline (abiotic effects) is important for determining variations in reproductive potential among lakes. Knowledge of these variations can be particularly useful in simulation modelling. Of great importance is also describing how $R F$ changes within a population. These variations are due to density-dependent factors (biotic effects) which affect growth, survival, abundance and fecundity. Within-lake studies describing temporal changes in walleye fecundity are scarce. Baccante and Reid (1988) describe changes in fecundity in Savanne and Henderson lakes walleye over a six-year period. They hypothesized that variations in food abundance within each lake, were responsi-

Table 4. Estimated abundance (number of fish), density (number of fish per hectare), and survival rates (percent) from egg stage to each age groups for walleye in Savanne Lake (Ontario), and Escabana Lake ${ }^{(1}$ (Wisconsin).


[^0]ble for changes in fecundity during the study period.
More recently, Muth and Ickes (1993) compared 1966 walleye fecundity (Wolfert 1969) to 1990-91 data. They found that, mean egg production of the
dominant age group of spawners (ages 4 to 8 ) was approximately $25 \%$ lower in 1990-91 than 1966. However, $R F$ was much higher in 1990-91, 84710 eggs per kilogram (K. Muth, Sandusky, OH pers.

Table 5. Morphometric, yield and egg production data for walleye populations in eleven lakes. Units as indicated. Yields in kg/ha, densities in number/ha and egg data in thousands.

|  | Wapun | Home | Savanne | Mille Lacs | Escanaba | Otter Tail |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Location | MB | MB | ON | MN | WI | MN |
| Degree days | 1200 | 1200 | 1367 | 1905 | 1901 | 1910 |
| Area (ha) | 213 | 170 | 364 | 53419 | 119 | 5968 |
| Shoreline (km) | 18 | 11 | 15 | 126 | 8 | 14 |
| Maximum depth (m) | 7 | 4 | 4 | 11 | 8 | 38 |
| Mean depth (m) | 3 | 2 | 2 | 6 | 4 | 13 |
| TDS (mg/l) | 45 | 50 | 42 | 133 | 34 | 228 |
| MEI | 14 | 28 | 16 | 22 | 8 | 19 |
| Total potential yield | 4.6 | 6.3 | 4.9 | 5.6 | 3.6 | 5.2 |
| Walleye yield | 3.4 | 1.6 | 1.8 | 3.0 | 7.0 | 2.8 |
| Exploitation rate (\%) | 52 | 23 | 16 | 25 | 28 | 25 |
| Adult density | $15^{\text {a) }}$ | $14^{\text {a) }}$ | $15^{\text {b) }}$ | 30 | $41^{\text {c) }}$ | 28 |
| Fingerling density |  |  | $28^{\text {d) }}$ |  | $59^{\text {d) }}$ |  |
| Standing crop | 10 | 10 | 7 | 14 | 19 | 12 |
| Inst. mortality | 0.8 | 0.4 | 0.7 | 0.5 | 0.5 | 0.7 |
| Annual mortality | 0.5 | 0.3 | 0.5 | 0.4 | 0.5 | 0.5 |
| Mean eggs/kg | 52.5 | 40.5 | 39.7 |  | 67.9 | 45.3 |
| Range eggs/kg | 24-61 | 34-51 | 29-67 |  | 48-74 |  |
| Pot. egg prod. ( $\times 10^{6}$ ) | 55.2 | 34.7 | 83.8 | 23 247 ${ }^{\text {e }}$ | 85.9 | 1596 |
| Eggs/ha | 260 | 210 | 229 | 435 | 724 | 268 |
| Eggs/kg of yield | 76 | 131 | 128 | 145 | 103 | 96 |
| Eggs: standing crop | 0.006 | 0.007 | 0.007 | 0.007 | 0.006 | 0.010 |
| Data source | 1 | 1 | 2 | 3 | 4 | 5 |
|  | Big Sand | Lake of the Woods | Leech | Dauphin | St. Clair |  |
| Location | MN | MN | MN | MB | ON |  |
| Degree days | 1900 | 1657 | 1900 | 1500 | 2500 |  |
| Area (ha) | 671 | 125858 | 45123 | 52200 | 75600 |  |
| Shoreline (km) | 13 |  | 314 | 119 | 270 |  |
| Maximum depth (m) | 41 | 12 | 46 | 4 | 6 |  |
| Mean depth (m) | 14 | 7 | 6 | 3 | 3 |  |
| TDS (mg/l) | 190 | 138 | 172 | 353 | 123 |  |
| MEI | 14 | 19 | 31 | 141 | 41 |  |
| Total potential yield | 4.6 | 5.3 | 6.5 | 13 | 7.4 |  |
| Walleye yield | 2.3 | 1.6 | 2.1 | 0.3 | 0.9 |  |
| Exploitation rate (\%) | 20 | 15 | 25 |  | 8 |  |
| Adult density | 19 | 8 | 14 |  | 8 |  |
| Fingerling density |  |  | 83 |  |  |  |
| Standing crop | 24 |  |  |  |  |  |
| Inst. mortality |  |  |  |  | 0.4 |  |
| Annual mortality |  |  |  |  | 0.3 |  |
| Mean eggs/kg | 61.4 | 60.3 |  |  | 87.4 |  |
| Range eggs/kg |  |  |  |  | 59-233 |  |
| Pot. egg production | 409 | 16400 | $11000{ }^{\text {e }}$ | 2817 | 5650 |  |
| Eggs/ha | 609 | 130 | 244 | 54 | 75 |  |
| Eggs/kg of yield | 264 | 80 | 116 | 217 | 84 |  |
| Eggs:standing crop | 0.003 | 0.006 | 0.006 |  | 0.01 |  |
| Data source | 6 | 3 | 3 | 1 | 7 |  |



1. J. Babaluk (Fisheries \& Oceans, MB, pers. comm.), 2. Our study, 3. D. Schupp (MN DNR, pers. comm.), 4. S. Serns (WI DNR, pers. comm.) and Serns (1982), 5. D. Schreiner (MN DNR, pers. comm.), 6. P. Jacobson (MN DNR, pers. comm.), 7. D. MacLennan (OMNR, pers. comm.).
comm.) versus 61149 eggs per kilogram in 1966 (Wolfert 1969). The increase in $R F$ is due to a decrease in mean weight-at-age from 1966 to 1990-91.

Colby and Nepszy (1981), and more recently Henderson and Nepszy (1994), have shown how environmental factors influence maturity. Northern stocks mature later and over a greater number of years than southern stocks. Rate of maturity is affected by growth, thus faster-growing stocks will reach maturity earlier. As a result, more southern walleye populations have a greater capacity to compensate to exploitation than northern ones. However, in the extreme southern range of their distribution, the lack of sufficient cool temperatures may inhibit egg development (Colby \& Nepszy 1981).

### 4.4. Exploitation rates

As our results indicate, the frequency distribution of exploitation rates is skewed to the right. This reflects the fact that most of the fisheries for which we have exploitation rates data, are from medium to heavilyexploited populations in the upper U.S. Midwest. Very little information is available for lakes with low fishing intensity, representative of the large number of lakes in northern Ontario. The reason for this lack of data is that, in northern Ontario creel surveys to estimate harvest rates are usually initiated after a decline in fishing success and yield was perceived (Baccante \& Colby 1991).

Spangler et al. (1977) provide a synthesis of response indicators of percids to exploitation. They conclude that changes in recruitment variability, growth rates and age to maturity are the most conspicuous responses to exploitation. Thus, the maximum sustainable exploitation rate depends on how rapidly the population can compensate to reductions in density within the bounds of habitat and energy availability. Although exploitation rates reported in the literature may not reflect maximum sustainable rates for the respective fisheries, they likely reflect rates characteristic of fairly heavily exploited fisheries because of the demand for walleye.

We plotted walleye exploitation rates over an energy cline, as measured by $G D D$ (Fig. 8). Although we cannot quantify the sustainability of these rates, using few, long-term data sources (Savanne Lake, Ont.; Wisconsin lakes and Lake Erie), we drew a line which we feel approximates levels beyond which
over exploitation is likely to occur. It is difficult to generalize about allowable exploitation rates because they are influenced by factors such as lake productivity, habitat and management objectives, however, we feel that very few walleye populations will sustain exploitation rates beyond $30 \%$, without significant loss in fishing quality. A lake with lower productivity will not support as high exploitation rates as a more productive lake within the same $G D D$ zone. For example, Whitefish and Savanne lakes, Ontario, both receive about $1900 G D D$, however Whitefish is much more productive than Savanne, $M E I$ of 33 and 16, respectively. Higher productivity translates into higher adult population density, about 25/ha in Whitefish and 15/ha in Savanne. Habitat can also influence survival, density and ultimately allowable harvest. Deep, clear lakes favour lake trout/ bass/pike communities with marginal walleye populations which cannot support harvest rates similar to populations in shallower, dark water lakes with more favourable walleye habitat (Johnson et al. 1977). Management objectives also dictate the rate of exploitation. If the management objective is to preserve fishing quality, then harvest should be kept at a much lower level than the maximum allowable rate. However, if the objective is to maximize angling harvest opportunities, then higher fishing quality may be traded off for higher harvest, set within the biological constraints of the system (Baccante \& Colby 1991).

Exploitation rates that have resulted in collapses of walleye populations in four lakes are also shown on the graph (Fig. 8). The waterbodies are: Henderson Lake (Ontario), Nipigon Bay (Lake Superior, Ontario), Wapun Lake (Manitoba) and Wolf Lake (Alberta). Henderson and Wapun Lakes fisheries were exploited as part of research studies by the Ontario Ministry of Natural Resources and Fisheries and Oceans Canada, respectively. The exploitation rates in both lakes were very high, around the $50 \%$ range (Table 3). The walleye in Nipigon Bay was mostly harvested by commercial fishery, however, Ryder (1968) postulated that pollution from a kraft paper mill was also responsible for the demise of the population. Although fishing mortality increased from 7 to $34 \%$ from 1955 and 1957, Ryder felt that this was of secondary importance compared to habitat degradation. The walleye population in Wolf Lake, Alberta, has been subjected to tremendous increases in angling pressure. From 1979, when
creel surveys were first started, to 1992 , angling pressure has increased $600 \%$ (Mike Sullivan, Dept. of Environ. Protection, Edmonton, Alberta, pers. comm.). Sullivan also reports significant increases in angler skill levels, and better fishing equipment (sounders, Lindy Rigs, leeches, etc.). As the walleye density declined in Wolf Lake, the fish concentrated in two small areas, which made them extremely vulnerable to anglers.

Although exploitation rates as high as 35\% are considered sustainable in some lakes, for example in Wisconsin (Staggs et al. 1990), we feel that exploitation rates exceeding $25 \%$ is probably optimum in high energy systems with good walleye production. One lake which has produced unusually high walleye yields ( $9 \mathrm{~kg} / \mathrm{ha}$ ) over a long time period (over 30 years) with exploitation rates averaging around $28 \%$, is Escanaba Lake, Wisconsin. Escanaba has relatively low fertility (MEI of 7.9) but good survival of young walleye which grow fast, mature early and have high fecundity. We suspect that Escanaba's combination of dark water, which provides competitive advantage for walleye, and sufficient weed beds to protect young fish from cannibalism and predation, provides ideal conditions for high turnover rates. Also, yellow perch do not grow to large sizes, possibly due to predator density, thus reducing the impact on young walleye. Overall, energy in Escanaba is efficiently channelled through the walleye component of the community.

In less productive systems, typical of a large number of boreal lakes in northern Ontario and other parts of Canada, it is unlikely that many populations can support exploitation rates in excess of around $15 \%$, without significant loss of fishing quality. However, as we have already pointed out, the scarcity of exploitation rates data for these types of lakes makes it difficult to predict safe rates over a broad geographic area. However, at the northern limits of their distribution, even exploitation rates of $10 \%$ may be too high for adult walleye. Confidence in setting safe rates is enhanced if managers have available adequate population, community and habitat information.

## 5. Management implications

Comparing population variables for walleye throughout their distributional range provides insights regarding the magnitude of variation among these
parameters. For instance, by tabulating the data into quartiles (Table 1), we describe quantitatively where a given body of water might lie along a gradient of population density, exploitation rates or yield. This comparative approach helps to categorize a lake based on its population characteristics, and helps to derive appropriate management actions.

A comparative approach is useful because long term data sets are scarce. More common are single year observations or sporadic ones, often a few years apart. This is true particularly for estimates of population abundance. Hansen et al. (1991) used longterm data from Escanaba Lake, Wisconsin, to evaluate the accuracy of population estimates from other lakes where reliable data were lacking. The estimates were then used to set harvest quotas. They concluded that the accuracy of estimates of adult walleye abundance declined over time from the year the estimates was obtained to the year it was used to set a harvest quota. This underscores the need and importance of long-term studies as reference points when comparing single year or a limited number of observations. Consider a given walleye population having an average density, and a higher than average exploitation rate. We would than expect the average size of fish harvested to be smaller, and the sustainable harvest to be average or better. In more northern latitudes especially on the Canadian shield we expect to find lower sustainable exploitation rates and yields associated with declining insolation and nutrient availability, a sustainable harvest of larger size fishes (due to slower growth and delayed maturity), more variable densities due to slower recovery rates.

If, observations on a remote fishery appear aberrant or unexpected, in respect to quartiles (Table 1) an investigator might consider the data suspect , and request further research or information initiating regulations. For example, population estimates may be suspect if other parameters such as harvest, exploitation rates, harvest size, and potential yields appear incompatible with other population parameters. The same would apply when comparing estimates of the other variables with each other. Finally, we suggest that a person trained in monitoring the status of walleye populations will find the comparative information presented here useful for interpreting harvest and environmental impacts on lightly monitored stocks.

[^1]
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