

Temporal and spatial variation in metric asymmetry in skulls of polar bears (*Ursus maritimus*) from East Greenland and Svalbard

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Fluctuating asymmetry (FA) as a measure of environmental stress was studied in polar bear (*Ursus maritimus*) skulls from East Greenland ($n = 300$, collected 1892–2004) and Svalbard ($n = 391$, collected 1950–2004). Nine metric traits in skull and lower jaw were measured. FA levels of each trait were compared between sex/age groups (subadults, adult females, adult males), periods (≤ 1960 , > 1960), and localities (East Greenland, Svalbard). The period ≤ 1960 was chosen to represent a period prior to the appearance of organohalogen pollution in the Arctic. Results indicated that Svalbard bears generally had a higher level of FA than those of East Greenland. Overall, no substantial evidence of a linkage between FA and organohalogens was found. Instead, indications were of subpopulations with declining levels of FA over time, suggesting the existence of positive population level effects powerful enough to overrule the influence of stress caused by global warming, pollution, and overharvesting.

Introduction

Fluctuating asymmetry (FA) refers to small, random deviations from the ideal morphological symmetry and is typically measured as the absolute difference between a trait on the left- and the right-hand side of a bilaterally symmetrical organism (Møller & Swaddle 1997, Palmer & Strobeck 2003a). Generally, one must presume that the two sides of the same individual develop

according to the same genetic programme and are exposed to identical influences from the extrinsic environmental factors, such as temperature, noise, food access and quality, chemicals, parasites, and diseases (e.g. Siegel & Mooney 1987, Borisov *et al.* 1997, Stige 2004, Jones *et al.* 2005, Møller 2006, Petavy *et al.* 2006). FA is often taken as a measure of developmental stability (DS), “the ability to attain equal development under the given circumstances” (Zakharov

1992), and its counterpart developmental instability (DI). In principle, DS reflects the organism's ability to buffer its development against disturbance (Møller & Swaddle 1997). The more fit the genotype of an organism, and the less stress from the surrounding environment it is exposed to, the higher DS and thus lower FA it is expected to have (Stige 2004). FA is therefore often taken as an indicator of phenotypic and genotypic quality (Møller & Swaddle 1997). As such, it can be used as a helpful tool in population dynamics research, where FA measurements could potentially be used as a substitute for, or as a supplement to, the more traditional ways of estimating the condition and fitness of a population (Clarke 1995, Baranov *et al.* 1997, Zakharov *et al.* 1997a, Stige 2004). High levels of FA have been connected with inbreeding depression (Baranov *et al.* 1997), high population density (Zakharov *et al.* 1997b), and genetic factors such as increased levels of hybridization and heterozygosity (Borrell *et al.* 2004, Andersen *et al.* 2006). Studies of FA have been conducted on metric and meristic traits of a multitude of organisms (plants: Pelabon *et al.* 2006, arthropods: Lajus & Alekseev 2004, Vilisics *et al.* 2005, Andersen *et al.* 2006, fish: Jagoe & Haines 1985, Green & Lochmann 2006, mammals; Zakharov & Yablokov 1990, Pertoldi *et al.* 2003).

An organism's DS is often believed to be largely determined during the foetal stage (e.g. Siegel *et al.* 1977b, Møller & Swaddle 1997, Valetsky *et al.* 1997). However, aspects such as canalization (Nijhout & Davidowitz 2003) and plasticity (Nijhout & Davidowitz 2003, Palmer & Strobeck 2003a) are also important concerns. A robust methodology is essential in FA studies to avoid serious flaws in the analyses (e.g. Merilä & Björklund 1995, Stige 2004). One of the greatest challenges in working with FA is getting the measurement error (ME) as small as possible (e.g. Merilä & Björklund 1995, Palmer & Strobeck 2003a, Stige 2004). In order to quantify ME in relation to FA it is important to take replicate and independent measurements (Palmer 1994, Merilä & Björklund 1995, Palmer & Strobeck 2003a). The obtained measurements must also be examined for directional asymmetry (DA) and anti-symmetry, as such measurements are presumed to be a result of other factors than a

low DS (Palmer & Strobeck 2003a, Stige 2004).

Many studies have been conducted in order to identify which forms of stress can affect FA and DS. Factors such as social stress (Valetsky *et al.* 1997, Gibbs & Breuker 2006), inbreeding (Pertoldi *et al.* 2000a), sexual selection (Voigt *et al.* 2005), nutritional stress (Pravosudov & Kitaysky 2006), heat stress (Siegel *et al.* 1977a, Petavy *et al.* 2006), and disease and parasitic stress (Møller 2006) have been found to be positively correlated with levels of FA. Organic pollution is one form of environmental stress that has been another positively correlated with FA (e.g. Zakharov & Yablokov 1990, Borisov *et al.* 1997, Zakharov *et al.* 1997a, Maul & Farris 2005). The Arctic acts as a sink for a range of persistent organic pollutants (POPs), most of them organohalogenes (OHCs) such as polychlorinated biphenyls (PCBs), chlordane (CHL), dichlorodiphenyltrichloroethane (DDT), dieldrin, hexachlorocyclohexane (HCH), hexachlorobenzene (HCB), and polybrominated diphenyl ethers (PBDEs) (AMAP 2004). Many of these pollutants are highly lipophilic and bio-accumulate in marine food chains, particularly in the fatty tissue of marine mammals (e.g. Borgå *et al.* 2001, AMAP 2004, Muir *et al.* 2006). Because the polar bear is an Arctic top predator, whose diet mainly consists of blubber from ringed seals (*Phoca hispida*) and bearded seals (*Erignathus barbatus*) (Ramsay & Stirling 1988, Derocher *et al.* 2002), it has indeed been found to carry some of the highest POP loads of any Arctic mammal species, especially so in bears from Svalbard and East Greenland (Norstrom *et al.* 1998, Dietz *et al.* 2004, Verreault *et al.* 2005, Muir *et al.* 2006, Dietz *et al.* 2007, Gebbink *et al.* 2008). Chronic exposure to these pollutants in the polar bears has been associated with afflictions such as endocrine disruption, reduced size of sexual organs and bone mineral density, and organ histopathology (e.g. Wiig *et al.* 1998, Derocher *et al.* 2003, Braathen *et al.* 2004, Lie *et al.* 2004, Sonne *et al.* 2004, Kirkegaard *et al.* 2005, Sonne *et al.* 2005a, 2005b, Sonne *et al.* 2006a, 2006b).

The use and outputs of for example PCBs, one of the major organochlorines, began in the 1950s and peaked in the 1960s (AMAP 2004, Dietz *et al.* 2004). Reports on the temporal trends of the persistent organic pollutants in

the Arctic now show indications of a decline in several of the “older” toxic substances over the last decades (e.g. Derocher *et al.* 2003, Dietz *et al.* 2004, Verreault *et al.* 2005, Riget 2005), while many novel contaminants seem to increase (Corsolini *et al.* 2002, Bossi *et al.* 2005, Muir *et al.* 2006, Smithwick *et al.* 2006). The potential adverse health effects from this organohalogen cocktail are unknown and make for a complex pattern when attempting to link organohalogen pollutants with FA.

Sonne *et al.* (2005a) investigated FA in a sample of polar bears from East Greenland and found no relation between pollution loads and FA. This present study is different from the aforementioned one in that the measuring equipment used is electronic and more exact, meaning lower ME and thus, supposedly, more robust results. Also, all traits were measured twice, and more robust statistical methods were applied than those reported in Sonne *et al.* (2005a), to obtain a more robust estimation of ME. In addition, the present investigation included polar bears from the Svalbard subpopulation, thus enabling a more extensive and comparative study with the added geographical element. The overall aim of the present study was to study metric FA in a combined sample of skulls of East Greenland and Svalbard polar bears in order to obtain information about spatial and temporal trends in the effects of environmental stress on DS.

Materials and methods

Samples

A sample of 300 polar bear skulls from East Greenland (held at the Zoological Museum, University of Copenhagen), and 391 polar bear skulls from Svalbard (held at the Natural History Museum, University of Oslo) were examined (Table 1). The skulls were collected in 1892–2004 (East Greenland, app. 61°–82°N, 10°–42°W) and 1950–2004 (Svalbard, app. 74°–81°N, 10°–35°E). Many of the skulls were more or less damaged, and thus the entire range of measurements could not be performed on all skulls. The age determination of the skulls had been done by counting the cementum Growth

Layer Groups (GLG) of the lower right incisor (I3) after decalcification, thin sectioning (14 Am) and staining (Toluidine Blue), using the method described by Hensel and Sorensen (1980) and Dietz *et al.* (1991). The age of 271 skulls from East Greenland and 227 skulls from Svalbard had been determined. Individuals were categorised as subadults (S), adult females (F) and adult males (M) as follows: Adult males ≥ 6 years old, adult females ≥ 5 years old and the remaining as subadults (Rosing-Asvid *et al.* 2002). Skulls were classified into the periods ≤ 1960 and > 1960 according to year of birth (age subtracted from year of kill). Measurements of individual levels of PCBs, HCB, HCHs, DDTs, dieldrin, CHLs, and PBDEs in subcutaneous adipose tissue were available from a subsample of 65 of the bears from East Greenland. Further details on the levels of the organohalogen compounds in relation to season, sex, and age are given in Dietz *et al.* (2004, 2007), Sandala *et al.* (2004), and Muir *et al.* (2006).

Table 1. Number of skulls of polar bears (*Ursus maritimus*) from East Greenland and Svalbard used in analyses of fluctuating asymmetry of the present study.

	East Greenland	Svalbard
Bears measured in total	300	391
Subadult	133	131
	(68 ♀, 58 ♂)	(51 ♀, 65 ♂)
Adult female (≥ 5 years)	59	89
Adult male (≥ 6 years)	73	111
Bears of known age (and hence year of birth)		
Subadult	133	98
Adult female	59	60
Adult male	73	63
Bears born ≤ 1960		
Subadult	40	9
Adult female	24	43
Adult male	49	58
Bears born > 1960		
Subadult	93	101
Adult female	35	10
Adult male	24	13
Bears born ≥ 1950		
Subadult	98	89
Adult female	44	49
Adult male	36	55
Bears with available contaminant data and age		
Subadult	36	0
Adult female	18	0
Adult male	11	0

Measurements

Nine metric bilateral traits were measured in order to estimate the level of FA in the skulls (Table 2 and Fig. 1). Each trait was measured on the right (*R*) and left (*L*) side of each skull, using digital callipers (Mitutoyo, Mitutoyo Corporation, Japan) to the nearest 0.04 mm. The traits were measured twice on each skull, always by the same person, and never twice on the same day.

Identification of sex

The Svalbard subadult sample consisted of 131 individuals, of which 90 were of known sex (52 male, 38 female). The adult sample consisted of 235 individuals, of which 137 were of known sex (84 male, 53 female). In the East Greenland sample, 88 of the 133 subadults were of known sex (47 male, 41 female), and 132 of the 133 adults were of known sex (73 male, 59 female). Classificatory analyses (*see below*) were run on the three groups with a larger number of individuals of undetermined sex (Svalbard: Subadult and adult, East Greenland: Subadult) (Hair *et al.* 2006). This was done in order to reach a higher number of individuals (*n*) with known sex and thus more degrees of freedom (*df*) in later analyses. The first of the repeated measurements (right side of all traits) were used as depend-

ent variables when assessing normality for the bears of known sex in each subgroup. Males and females were handled separately. Normality was assessed by applying the Shapiro-Wilk test solely to the bears of known sex with valid values in all dependent variable categories, as these were the bears which would be used in the further classificatory analyses. Factor analysis with a principal components extraction, stepwise binary logistic regression, and stepwise discriminant analyses were run in order to obtain suggestions for different models to discriminate between males and females. These analyses helped in understanding which variables (traits) were of greater importance when looking to discriminate between the sexes. The variables (traits) of least importance could thus be sorted out, and the analyses run again. The fewer the variables applied in each analysis, the higher the number of individuals (*n*) included in the calculation of the model would be. The higher *n* would cause the remaining variables to better meet the normality distribution.

Simultaneous discriminant analyses were used to analyze the various suggestions for a model, and the best solution for this study (a combination of high percentage of individuals with known sex correctly identified and a high number of individuals with unknown sex identified) was chosen. The resulting classifications were applied to the data set and used in further analyses.

Table 2. Definition of the nine metric traits measured on skulls of East Greenland and Svalbard polar bears (*Ursus maritimus*).

Trait	Definition
Skull	
OPF	Maximal distance between the opisthocranium and the postorbital process of the frontal bone.
CBL	Condylbasal length. The maximal distance between the anterior margin of the alveole of the 1st incisor and the anterior margin of the occipital condyles.
P4–M2	Length from the anterior margin of the 4th premolar to the posterior margin of the 2nd molar.
C–M2	Length from the anterior of the alveole of the canine, to the posterior of the alveole of the 2nd molar.
I1–M2	Length from the anterior of the alveole of the 1st incisor, to the posterior of the alveole of the 2nd molar.
POH	Postorbital height. The minimal distance between the postorbital process of the frontal bone and the frontal process of the zygomatic arch.
Lower jaw	
Lp4	Length of the 4th premolar of the lower jaw.
ML	Mandible length. The maximal distance between the anterior margin of the mandibular symphysis and the posterior margin of the angular process.
MH	Mandible height. The maximal distance between the margin of the angular process and the coronoid process.

Statistical analyses

Preliminary examination of data

All traits were tested with regards to normality, anti-symmetry, size dependency, and directional asymmetry (DA). Traits OPF, C–M2, I1–M2, POH and Lp4 all exhibited ideal FA. Significant DA was discovered in two traits, P4–M2 ($p < 0.0001$) and ML ($p < 0.0001$). If deviations from symmetry are heritable in traits that exhibit DA, then apparent differences in DS among groups could arise solely due to differences in the amount of inherited variation for asymmetry (Palmer 1994). In other words, DA goes against the theoretic “common ground” of FA, the ability of individuals in a given group to attain equal development under the given circumstances. In order to check the influence of DA on FA, analyses on traits P4–M2 and ML were also carried out using the absolute value of the residuals of a regression of the left and right side traits instead of simply $(lR - L)$ as the dependent variable (Pertoldi *et al.* 2000b). Only two results changed when using the residuals instead of the general FA-index: In the three-way ANOVA estimating FA within periods, locations and sex/age groups (*see below*), $ML_{(Resid.)}$ was not significant in any interactions (Table 3), but in the main factor sex/age group ($p = 0.008$). Furthermore, the significant result for trait P4–M2 in ΣPCB (Table 4), changed to non-significant when using $P4-M2_{(Resid.)}$. As the changes were so few and of no great consequence to the discussion, we decided to report the figures here, but not to change the method used throughout the paper.

Traits CBL and MH were size dependent until successfully transformed, using $[lR - L]/((R - L)/2)$ (Palmer & Strobeck 2003a). For all further analyses, traits CBL and MH were used in their transformed form, while all other traits were used untransformed. None of the nine traits showed significant anti-symmetry. *See* Bechshøft *et al.* (2008) for further information on calculations of asymmetry, measurement error, and repeatability.

All statistical analyses were performed using SPSS (version 13.0) (SPSS Inc., Chicago, Illinois). The significance level was set to $p = 0.05$ unless other is specified.

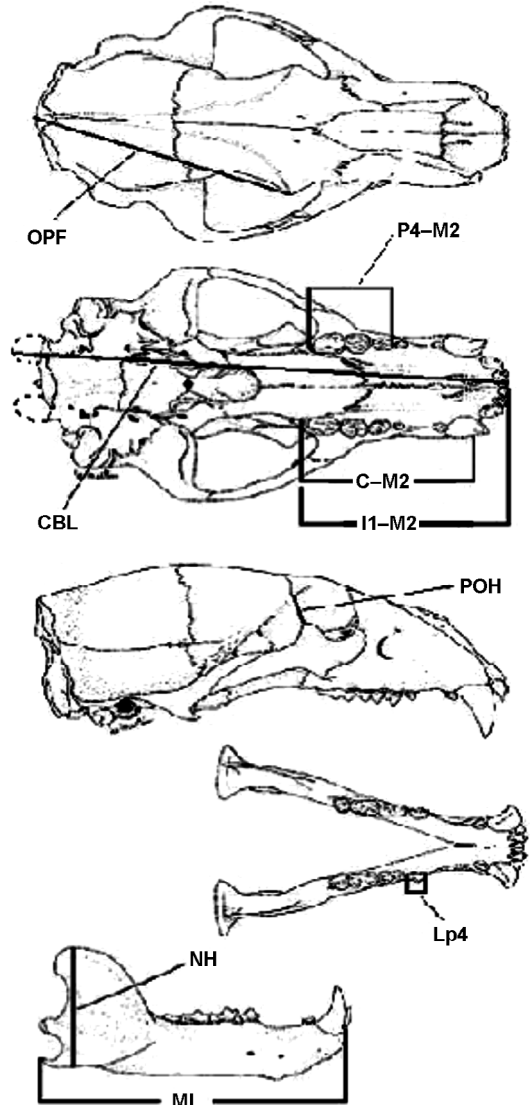


Fig. 1. Nine metric traits (*see* Table 2) measured on skulls of East Greenland and Svalbard polar bears (*Ursus maritimus*). Figure modified from Amstrup and DeMaster (1988).

Bias

FA may increase with increasing trait size (Leung 1998) and ME may increase with decreasing trait size (Merilä & Björklund 1995). This could create bias in the sample if the skulls of bears of any one sex or locality were consistently found to be growing larger or smaller over the period of time during which the investigated polar bear

Table 3. Significant factors and interactions from three-way ANOVAs (dependent variable: The absolute value of measured fluctuating asymmetry; independent variables: Locality, sex/age groups, periods) carried out on nine metric traits measured on skulls of East Greenland and Svalbard polar bears (*Ursus maritimus*), in order to estimate which variables play an important role in the differences observed in the measured fluctuating asymmetry.

Factors and interactions with a significant contribution		<i>p</i>
OPF	Adult male/adult female/subadults	< 0.0001
CBL	None	
P4–M2	(Adult male/adult female/subadults) × (≤ 1960 and > 1960)	0.042
C–M2	None	
I1–M2	Geographic locality	0.005
POH	Adult male/adult female/subadults	0.020
Lp4	None	
ML	≤ 1960 and > 1960	< 0.0001
	(Geographic locality) × (≤ 1960 and > 1960)	0.016
ML _(Resid.)	Adult male/adult female/subadults	0.008
MH	(Adult male/adult female/subadults) × (≤ 1960 and > 1960)	0.015

Table 4. Results from hierarchical multiple regression analyses (dependent variable: The absolute value of measured fluctuating asymmetry; explanatory variables: Contaminant levels in individual subcutaneous adipose tissue samples) carried out on nine metric traits measured on skulls of East Greenland polar bears (*Ursus maritimus*), in order to assess the degree to which the measured contaminant levels were able to predict the differences in the same individuals' measured levels of fluctuating asymmetry. Measurements of individual levels of PCBs, HCB, HCHs, DDTs, dieldrin, CHLs, and PBDEs in subcutaneous adipose tissue were available from a subsample of 65 of the East Greenland polar bears (further details on the levels of the organohalogen compounds in relation to season, sex, and age are given in Dietz *et al.* (2004, 2007), Sandala *et al.* (2004), and Muir *et al.* (2006)).

		Trait								
		OPF	CBL	P4–M2	C–M2	I1–M2	POH	Lp4	ML	MH
ΔR^2 ^(a)	M1 ^(a)	1.5	2.2	1.2	3.4	7.7	7.3	3.0	0.8	10.0
	M2 ^(a)	25.2	19.5	13.6	9.6	23.9	24.0	6.1	8.1	16.0
	M total ^(a)	26.7	21.7	14.7	13.0	31.7	31.3	9.2	8.9	26.0
ΣPCB	<i>p</i>	0.418	0.037*	0.851	0.416	0.656	0.007*	0.811	0.877	0.381
	<i>B</i> ^(b)	< 0.0001	< 0.0001	< 0.0001	< 0.0001	< 0.0001	< 0.0001	< 0.0001	< 0.0001	< 0.0001
HCB	<i>p</i>	0.435	0.709	0.044*	0.136	0.001*	0.647	0.770	0.389	0.113
	<i>B</i> ^(b)	–0.003	< 0.0001	0.004	0.004	0.006	0.001	< 0.0001	–0.003	< 0.0001
ΣHCH	<i>p</i>	0.147	0.985	0.974	0.265	0.867	0.003*	0.997	0.797	0.755
	<i>B</i> ^(b)	–0.004	< 0.0001	< 0.0001	–0.002	< 0.0001	–0.005	< 0.0001	0.001	< 0.0001
ΣDDT	<i>p</i>	0.004*	0.475	0.425	0.835	0.911	0.317	0.977	0.592	0.210
	<i>B</i> ^(b)	0.003	< 0.0001	< 0.0001	< 0.0001	< 0.0001	< 0.0001	< 0.0001	< 0.0001	< 0.0001
Dieldrin	<i>p</i>	0.736	0.517	0.344	0.582	0.282	0.082	0.962	0.146	0.266
	<i>B</i> ^(b)	0.002	< 0.0001	0.003	0.002	0.003	0.004	< 0.0001	0.006	< 0.0001
ΣCHL	<i>p</i>	0.977	0.177	0.146	0.163	0.177	0.022*	0.581	0.258	0.231
	<i>B</i> ^(b)	< 0.0001	< 0.0001	–0.001	–0.001	< 0.0001	–0.001	< 0.0001	–0.001	< 0.0001
ΣPBDE	<i>p</i>	0.189	0.157	0.181	0.965	0.996	0.651	0.147	0.838	0.688
	<i>B</i> ^(b)	0.007	< 0.0001	–0.004	< 0.0001	< 0.0001	–0.001	0.001	0.001	< 0.0001

* Statistically significant at $p = 0.05$.

^{a)} The ΔR^2 is given by: M1; number of percent of the measured variation in absolute FA explained by age and sex of the individuals. M2; additional number of percent of the measured variation in absolute FA explained by the regression model, when levels of contaminants are brought in as explanatory variables as well. M total = total number of percent of the measured variation in absolute FA explained by the regression model.

^{b)} The unstandardized coefficients (*B*) denote the parameter estimates one would use when creating the equation calculated by the regression model for each individual trait.

skulls were collected. Condylbasal length was chosen to represent general size. Two three-way ANOVAs (one on the subadults and one on the adults) were applied to test for size bias over time, with condylbasal length (from the first of the repeated measurements) as $(CBL_{\text{right}} + CBL_{\text{left}})/2$. The first ANOVA was conducted with condylbasal length as the dependent variable, and locality, sex/age groups (here: subadults, divided into males and females), and periods (≤ 1960 and > 1960) as independent variables. The second ANOVA was conducted with condylbasal length as the dependent variable, and locality, sex/age groups (here: adults, divided into males and females), and periods (≤ 1960 and > 1960) as independent variables.

Outliers

Possible outliers, with regards to measurements, were identified from scatter plots as suggested by Palmer and Strobeck (2003b). The outliers were then tested and removed according to Grubbs' test (Grubbs 1969, Palmer & Strobeck 2003b).

FA in relation to sex/age groups, periods, and localities

A three-way ANOVA (dependent variable: FA ($IR - LI$), independent variables: locality, sex/age groups, and periods) was carried out for each trait in order to estimate which factors played an important role for differences in the observed FA. Prior to this, a one-way ANOVA was conducted on the traits in order to justify the pooling of the sexes in the subadult group (dependent variable: FA ($IR - LI$), factor: subadult males vs. subadult females), and the segregation of the sexes in the adult individuals (dependent variable: FA ($IR - LI$), factor: adult males vs. adult females). Games-Howell and Tukey's HSD post-hoc tests were applied in the analyses of two traits (OPF and POH). The time of the split between the two periods (≤ 1960 and > 1960) was chosen because outputs and use of for example PCBs, one of the major organochlorines, began being widely used in the 1950s and peaked in the 1960s (AMAP 2004, Dietz *et al.* 2004).

If an organism-wide level of FA exists, each trait would in theory add some information about it. As each trait ideally provides an independent measure of the organism's FA, combining the information from multiple traits ought to give a stronger analysis (Palmer & Strobeck 2003a). Therefore, FA data ($|\ln(R) - \ln(L)|$) for traits C-M2, I1-M2, POH, and Lp4 were pooled and run in a four-way ANOVA (dependent variable: $|\ln(R) - \ln(L)|$, pooled for the four aforementioned traits, fixed factors: locality, sex/age groups, periods, and trait). The log-transformation was applied in order to remove any scale effect among the four traits.

FA over time

The analyses of temporal variation in FA were run only on polar bears born in 1950 or later, as the individuals born before that time were too few to conduct continuous time-trend analyses on. The East Greenland and Svalbard samples were pooled in the analyses. The two samples were roughly equal as to number of males and females, as well as number of adult males, adult females, and subadults (*see* Table 1). Trait I1-M2 was not included in the analysis, as it showed a geographical difference (*see* Results). A scatterplot of the FA ($IR - LI$) of each of the remaining traits versus the birth year of the individuals was produced for each sex/age group. Linear regression lines (including 95% confidence intervals) and LOESS (locally weighted scatterplot smoothing) lines were added to each scatterplot. The LOESS line was applied to allow the data to reveal its own natural pattern (Cleveland 1979, Prince & Silva 2002, Vieth *et al.* 2003). As the LOESS line confirmed the linearity of the relationship between year of birth and FA for the nine traits, Pearson correlation analyses were performed for each sex/age group on each trait, assessing the degree to which the individuals' year of birth had a significant effect on the observed FA.

FA and contaminant levels

Hierarchical multiple regression analyses (FA (IR

– *L*) as dependent variable; sex and age in years, and contaminants as explanatory variables) were conducted in order to assess the degree to which individual contaminant levels were able to predict the differences in the individual levels of FA. The explanatory variables were entered in two steps; first sex and age in years (model 1, M1), then contaminants (model 2, M2). In this way, variables were assessed in terms of what they added to the prediction of the dependent variable.

Results

Identification of sex

For the Svalbard subadult sample, trait P4–M2 was the best parameter for distinguishing between males and females, with 86.7% correctly identified. An additional 27 subadult individuals were sexed according to this model. For the Svalbard adult sample, the model consisting of traits P4–M2 and I1–M2 was the best (91.0% accuracy), determining the sex of a high number of individuals of previously unknown sex ($n = 63$). This model was also the best for the East Greenland subadult sample, as it identified a high percentage of the bears of known sex (88.0% accuracy) and also determined the sex of a high number of individuals of previously unknown sex ($n = 38$).

The unstandardized canonical discriminant functions were estimated as seen below. Cutting scores were calculated as described in Hair *et al.* (2006).

Svalbard adults:

$$D = -25.11 + 0.20 \times \text{length of trait P4-M2}_{\text{right}} + 0.09 \times \text{length of trait I1-M2}_{\text{right}} \quad (1)$$

(Male group centroid: 1.08,
female group centroid: –1.56.
Cutting score –0.48).

Svalbard subadults:

$$D = -20.83 + 0.36 \times \text{length of trait P4-M2}_{\text{right}} \quad (2)$$

(Male group centroid: 0.77,
female group centroid: –0.96.
Cutting score –0.19).

East Greenland subadults:

$$D = -23.76 + 0.21 \times \text{length of trait P4-M2}_{\text{right}} + 0.09 \times \text{length of trait I1-M2}_{\text{right}} \quad (3)$$

(Male group centroid: 1.18,
female group centroid: –1.27.
Cutting score –0.09).

Bias

The three-way ANOVA testing for bias of condylobasal length over time in the subadult sample showed that CBL was larger during the early period (≤ 1960) ($p = 0.013$), especially for the Svalbard polar bears. There was no observed temporal bias in size for the adult sample.

Outliers

Of the outliers identified from the scatter plots and run through Grubbs' test (Grubbs 1969), a total of 126 measurements (*R* and *L*) were excluded from the statistical analysis, which corresponds to 2.0% of all measurements taken.

FA in relation to sex/age groups, periods, and localities

The one-way ANOVA of FA showed no significant differences between subadult males and subadult females in any traits ($0.101 \leq p \leq 0.997$), and the groups were thus pooled into a single group of subadults. The analyses of adult males *vs.* adult females revealed significant differences in regard to their FA in three of the nine traits (males having higher levels of FA), and the groups were therefore kept separate in the further FA calculations. There were significant results from the three-way ANOVA estimating FA within periods, locations, and sex/age groups (Table 3): The FA of traits OPF ($p < 0.0001$) and POH ($p = 0.020$) differed significantly among the sex/age groups. Games-Howell (for OPF) and Tukey's HSD (for POH) post-hoc tests showed that adult males had a significantly higher FA than subadults (OPF: $p < 0.0001$, POH: $p = 0.046$) and adult females (OPF: $p <$

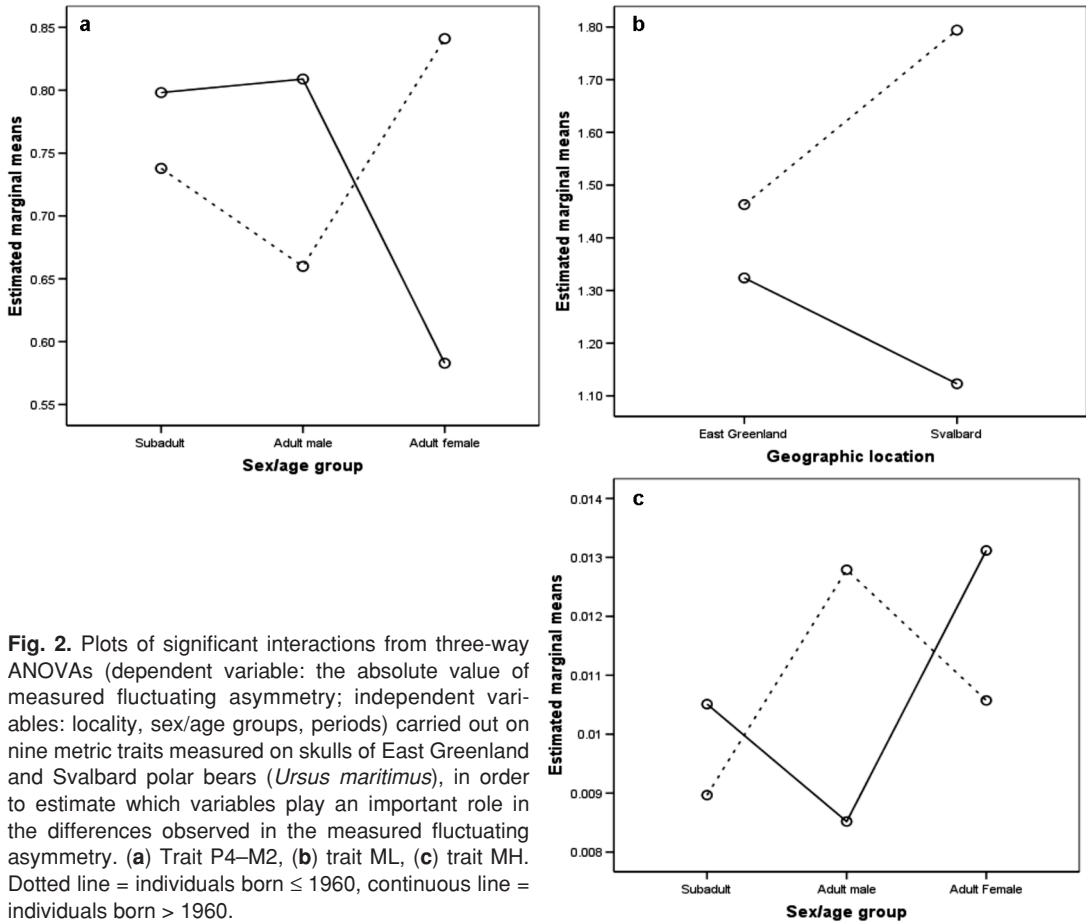


Fig. 2. Plots of significant interactions from three-way ANOVAs (dependent variable: the absolute value of measured fluctuating asymmetry; independent variables: locality, sex/age groups, periods) carried out on nine metric traits measured on skulls of East Greenland and Svalbard polar bears (*Ursus maritimus*), in order to estimate which variables play an important role in the differences observed in the measured fluctuating asymmetry. (a) Trait P4–M2, (b) trait ML, (c) trait MH. Dotted line = individuals born ≤ 1960 , continuous line = individuals born > 1960 .

0.0001, POH: $p = 0.036$). In traits P4–M2 and MH, the interaction between sex/age groups and periods was significant (P4–M2: $p = 0.042$; MH: $p = 0.015$). In trait P4–M2 subadults as well as adult males had a lower FA during the first period than during the second period (Fig. 2a). The pattern for trait P4–M2 of the adult females was the opposite, with a clearly lower FA during the second period. In the interaction in trait MH, adult males had a markedly lower FA during the second period, whereas the pattern for both subadults and females were the exact opposite, with highest FA-values in the second period (Fig. 2c).

The FA of trait I1–M2 was significantly higher ($p = 0.005$) in the Svalbard bears. The polar bears (irrespective of sample locality and sex/age group) born ≤ 1960 had a higher FA in trait ML than those born > 1960 ($p < 0.0001$).

Trait ML was also determined by the interaction effect between geographic locality and periods ($p = 0.016$); all polar bears from East Greenland as well as those from Svalbard had a higher FA during the early period (Fig. 2b). The difference between the two periods was by far the greatest in the Svalbard subpopulation (Fig. 2b).

The four-way ANOVA on the pooled four traits showed that trait ($p < 0.0001$) and locality ($p = 0.030$) were significant factors within the model. Trait POH held the highest FA followed by Lp4, C–M2, and finally I1–M2. All traits were significantly different from each other ($p < 0.0001$), except for C–M2 and I1–M2 ($p = 0.198$). The Svalbard polar bears had a significantly higher FA ($p = 0.030$) than those from East Greenland. All other factors, main as well as interactions, were non-significant.

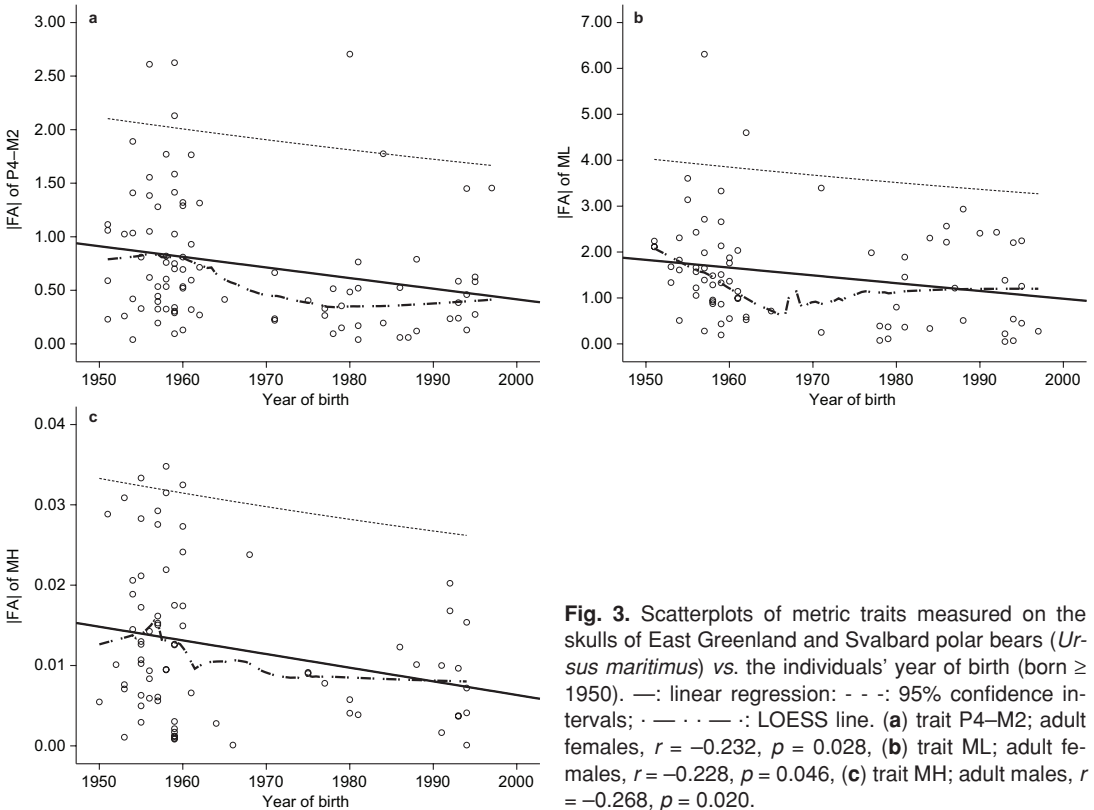


Fig. 3. Scatterplots of metric traits measured on the skulls of East Greenland and Svalbard polar bears (*Ursus maritimus*) vs. the individuals' year of birth (born \geq 1950). —: linear regression; - - -: 95% confidence intervals; · - · - ·: LOESS line. (a) trait P4–M2; adult females, $r = -0.232$, $p = 0.028$, (b) trait ML; adult females, $r = -0.228$, $p = 0.046$, (c) trait MH; adult males, $r = -0.268$, $p = 0.020$.

FA over time

For all traits, the linear regression and the LOESS lines largely followed the same path (Fig. 3), which rendered the linear correlation a plausible model for the nine traits in each subgroup (Cleveland 1979, Prince & Silva 2002, Vieth *et al.* 2003). There was a clear trend in the three significant results from the Pearson correlation analyses, as they all exhibited a negative correlation of FA over time: adult females, traits P4–M2 (Fig. 3a: $r = -0.232$, $p = 0.028$) and ML (Fig. 3b: $r = -0.228$, $p = 0.046$), and adult males, trait MH (Fig. 3c: $r = -0.268$, $p = 0.020$). There was no clear negative or positive tendency for the remainder of the results of the correlation analyses.

FA and contaminant levels

There were seven instances (out of $63 = 9$ traits \times 7 contaminant groups) of regression analyses where contaminant groups had a significant influ-

ence on the FA of individual traits; Σ DDT in trait OPF ($B = 0.003$, $p = 0.0004$), Σ PCB in trait CBL ($B < 0.0001$, $p = 0.037$), HCB in traits P4–M2 ($B = 0.004$, $p = 0.044$) and I1–M2 ($B = 0.006$, $p = 0.001$), and Σ PCB ($B < 0.0001$, $p = 0.007$), Σ HCH ($B = -0.005$, $p = 0.003$) and Σ CHL ($B = -0.001$, $p = 0.022$) in trait POH (Table 4). Sex and age of the polar bears explained between 0.8% and 10.0% of the variation in the measured FA (M1 in Table 4), whereas the contaminant levels explained between 6.10% and 25.2% of the variation in the measured FA (M2 in Table 4). When both factors (age and sex, and contaminant levels) were taken into account in the model, between 8.90% and 31.70% of the measured variation in FA was explained (M total in Table 4).

Discussion

Identification of sex

Classificatory analyses for identification of

sex were performed on Svalbard subadults and adults, and on East Greenland subadults. All models primarily included the trait P4–M2, which was also applied previously by Larsen (1971) in sex identification of polar bears.

Bias

FA may increase with increasing trait size (Leung 1998). This could create bias in the sample if the skulls were consistently found to be growing larger or smaller over the period during which the investigated polar bear skulls were collected. The three-way ANOVA assessing bias of condylobasal length of subadult polar bears over time showed that the CBL of the subadults was significantly longer during the first period (≤ 1960), especially for the Svalbard subadults. These results could be caused by the limited number of subadult Svalbard polar bears available from the earlier period compared to the second period. This bias must be taken into consideration when interpreting other results pertaining to the subadult bears in the present FA study.

Outliers

In the present study, 2.0% of the FA data were identified and treated as outliers. In comparison, Ahtiainen *et al.* (2003) identified 2.1% of their FA data as outliers, and of the data from Crespi and Vanderkist (1997), Palmer and Strobeck (2003b) identified 5.7% as outliers.

FA in relation to sex/age groups, periods, and localities

Trait ML of all sex/age groups had a FA which proved significantly higher during the first period than during the second. This obviously goes against the theory of pollution as generator of increasing FA. This could be due to the DA of the trait, or perhaps the change is instead caused by other extrinsic factors than pollution; e.g. population density, temperatures, or prey availability. Also, it could partly be a result of the previously mentioned bias in size over time found in the

subadult sample. The results for the interaction between traits P4–M2 and MH, respectively, and the two periods were rather inconsistent, again perhaps due to the DA of trait P4–M2. The only sex/age group that showed the same trend for both interactions was subadults; in traits P4–M2 and MH the subadults had a significantly higher FA during the second period than they did during the first. Perhaps the reason is that the subadult bears are under more severe stress during their phase of rapid growth (Derocher & Wiig 2002), with the fatty mother's milk they consume at this sensitive stage of their life containing high levels of organohalogenes (Bernhoft *et al.* 1997, Polischuk *et al.* 2002). An organism's DS is often believed to be largely determined during the foetal stage (Siegel *et al.* 1977b, Møller & Swaddle 1997, Valetsky *et al.* 1997). Still, the trait can apparently also be influenced to a certain extent throughout the organisms' entire lifespan. Møller and Swaddle (1997) suggest that the DS may depend on the nature of the trait, i.e. how functionally important it is to the organism. Vertebrate bones for example are capable of significant remodeling (Olsen *et al.* 2000). Therefore, aspects such as canalization (Nijhout & Davidowitz 2003) and plasticity (Nijhout & Davidowitz 2003, Palmer & Strobeck 2003a) are also important concerns.

Svalbard polar bears had significantly higher FA in trait I1–M2 than did the bears from East Greenland, a trend which was reflected in the outcome of the pooling of four of the traits. This gives reason to believe that the trend of Svalbard polar bears having higher FA levels than East Greenland polar bears could well be a trend in other traits too, although too faint to be detected when each trait is analyzed on its own. A geographical difference between the two subpopulations was also found by Sonne *et al.* (2007), with the Svalbard bears having a higher ratio of tooth wear than the East Greenland bears. Sonne *et al.* (2007) ascribed the observation to differences in bone consumption (food competition) caused by population density and food availability. Perhaps the difference in FA could also be ascribed to the speculated increased polar bear population found in Svalbard (Derocher 2005).

The adult males had significantly higher FA than both adult females and subadults in traits OPF and POH. The explanations for this lie in

speculations only; perhaps the males are generally under more stress than the other two sex/age groups, perhaps they differ in plasticity which could potentially affect their FA levels, or perhaps this could be a case of greater body size causing higher variance (Pertoldi *et al.* 2007). Should either of these hypotheses be true though, one would perhaps expect the adult males to stand out significantly from the other two sex/age groups in more than just two traits.

One could also speculate as to the choice of year of split between non-polluted and polluted periods. Perhaps a division into three periods could have been an option, as many of the “old” organochlorines have shown indications of decrease recently (Derocher *et al.* 2003, Dietz *et al.* 2004, Verreault *et al.* 2005, Riget 2005). Change in FA over time was therefore further analysed.

FA over time

All three significant results from the correlation analyses of FA over time (with the two subpopulations merged into one) were negatively sloped; traits P4–M2 and ML in the adult females, and trait MH in the adult males. The results for the females could be confounded by the DA found in traits P4–M2 and ML. Assuming that FA is an indicator of stress, the East Greenland and Svalbard polar bear subpopulations seem to have become progressively less stressed over time since the 1950s, in spite of possible effects of excessive harvesting (Lønø 1970, Larsen 1986, Wiig 2005), pollution (Norstrom *et al.* 1998, Skaare *et al.* 2000, Derocher *et al.* 2003, AMAP 2004, Braathen *et al.* 2004, Verreault *et al.* 2005, Muir *et al.* 2006), and a warming climate (ACIA 2005, Regehr *et al.* 2005, Wiig 2005).

If factors such as temperature or other extrinsic stressors have shifted considerably over the period during which the skulls were collected, it could explain why FA was found higher in the earlier period. Higher climatic fluctuations (temperature extremes) during the earlier period could have led to a decline in prey availability and thus starvation-induced stress and FA in the polar bears. However, this hypothesis is not likely, as temperatures above normal have been

experienced in the Arctic during the last decades (ACIA 2005, Sonne *et al.* 2005a). Earlier studies of marine mammals detected differences in FA over time, connected to pollution (Zakharov & Yablokov 1990, Schandorff 1997). However, in these studies, PCB and DDT concentrations were up to a hundred times higher than those found in the present study (*see e.g.* Olsson & Reutergårdh 1986, Blomkvist *et al.* 1992).

Apparently there must also exist other factors than the abovementioned to be reckoned with when assessing the state of these two polar bear subpopulations. It has recently been speculated that the Svalbard polar bear subpopulation, which was previously so heavily depleted (Lønø 1970, Larsen 1986), may now have reached a level of recovery where individual growth becomes density dependent (Derocher 2005). This would indicate that the subpopulation growth was currently at or beyond its peak. The present results, however, indicate that density dependence is not yet to be reckoned a stress factor for the East Greenland and Svalbard polar bear subpopulations. An alternative explanation could simply be that FA of metric skull traits is not a suitable predictor of stress in polar bears. However, metric skull traits have been successfully applied as measurers of stress in other species (Siegel *et al.* 1977b, Pertoldi *et al.* 2000a, Hoelzel *et al.* 2002).

A more speculative explanation for the decrease in FA found in the present study could be the declining pollution rates in most of the “older” organohalogenes (Derocher *et al.* 2003, Dietz *et al.* 2004). Concentrations of pollutants such as those examined in the present study have decreased ca. 20%–70% since 1990 (Dietz *et al.* 2004). Were these organohalogen concentrations extrapolated back in time, values observed around ca. 1965 would in all probability be in the same range as those used by Zakharov and Yablokov (1990) and Schandorff (1997). This would support our hypothesis of metric FA being linked to organohalogen pollutants in polar bears. However, no data is present to support this, and the existence of such a link thus remains uncertain.

FA and contaminant levels

The majority of the significant results from the

multiple regressions between FA and contaminant levels were positively correlated. However, parameter estimates (B) for all seven significant results were very low. This means that although a trend supporting the theory of increased pollution levels leading to more stress and higher FA was found, no clear picture emerged. None of the few other studies attempting this correlation between FA and contaminants on an individual level had particularly higher success in linking the two factors (Pertoldi *et al.* 1997, Hendrickx *et al.* 2003). The pollutants supposedly have their greatest effect on the individual at the foetal stage (e.g. Siegel *et al.* 1977b, Møller & Swaddle 1997, Valetsky *et al.* 1997). This is why the model was tested for influence by age. However, age and sex, even together, did not explain much of the variation found in the measured FA (M1, Table 4); the pollutants themselves explained a great deal more (M2, Table 4). As when analysing FA over time, one must probably take factors other than pollution into consideration when assessing the causes of fluctuating asymmetry. Perhaps FA of metric skull traits is not a suitable predictor of pollution-induced stress in polar bears. Or perhaps the organohalogen toxic loads are not at high enough levels to elicit a measurable positive response in the FA of the examined polar bears. Schandorff (1997), and Zakharov and Yablokov (1990) found that high levels of FA were associated with high levels of pollutants, but the pollutant concentrations in these studies by far exceeded the ones in the present study, as also mentioned earlier.

The present FA study was in part motivated by the results of Sonne *et al.* (2005a). The aim was to achieve better and more robust results when linking the FA of the bears to time, and to organohalogen loads in and around them. This was attempted by including a higher number of bears, and by applying more exact measuring equipment and methods. The measuring methods and statistical approach of the present study succeeded in being more robust; in particular in regards to the repeated measurements of all traits, which provided an excellent basis for comparing measurement errors and levels of FA (Bechshøft *et al.* 2008). However, this did not help create a markedly clearer picture than the one obtained in Sonne *et al.* (2005a) when linking the FA of the

polar bears to time and organohalogen loads. As the metric traits used in this study were measured, foramina were counted on the same polar bear skulls. Analysing these meristic data for FA is a natural next step, and may help clarify the picture.

Conclusions

The results indicated that Svalbard bears generally had a higher level of FA than those of East Greenland. Subadult polar bears were found to have higher FA in two traits during the late period (> 1960), while adult males were found to have generally higher FA in two traits than the other sex/age groups. Overall, the present study showed no substantial evidence of a linkage between levels of FA and organohalogenes. Instead, the indications were of subpopulations with generally declining levels of FA over time, suggesting the existence of positive population level effects powerful enough to overrule the negative influence of stress caused by global warming, pollution, and overharvesting.

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References

- ACIA 2005: *Arctic climate impact assessment*. — Cambridge University Press, Cambridge.
- Ahtiainen, J. J., Alatalo, R. V., Mappes, J. & Vertainen, L. 2003: Fluctuating asymmetry and sexual performance in the drumming wolf spider *Hygrolycosa rubrofasciata*. — *Annales Zoologici Fennici* 40: 281–292.
- AMAP 2004: *AMAP assessment 2002: Persistent organic pollutants in the Arctic*. — Arctic Monitoring and Assessment Programme, Oslo.
- Amstrup, S. C. & DeMaster, D. P. 1988: Polar bear (*Ursus maritimus*). Biology, management and conservation. — In: Lentfer, J. W. (ed.), *Selected marine mammals of Alaska: Species accounts with research and management recommendations*: 39–56. Marine Mammal Commission, Washington, DC.

- Andersen, D. H., Pertoldi, C., Loeschcke, V. & Scali, V. 2006: Developmental instability, hybridization and heterozygosity in stick insects of the genus *Bacillus* (Insecta, Phasmatodea) with different modes of reproduction. — *Biological Journal of the Linnean Society* 87: 249–259.
- Baranov, A. S., Pucek, Z., Kiseleva, E. G. & Zakharov, V. M. 1997: Developmental stability of skull morphology in European bison *Bison bonasus*. — *Acta Theriologica Suppl.* 4: 79–85.
- Bernhoft, A., Wiig, Ø. & Skaare, J. U. 1997: Organochlorines in polar bears (*Ursus maritimus*) at Svalbard. — *Environmental Pollution* 95: 159–175.
- Bechshøft, T. Ø., Rigét, F. F., Wiig, Ø. & Sonne, C. 2008: Fluctuating asymmetry in metric traits; a practical example of calculating asymmetry, measurement error, and repeatability. — *Annales Zoologici Fennici* 45: 32–38.
- Blomkvist, G., Roos, A., Jensen, S., Bignert, A. & Olsson, M. 1992: Concentrations of DDTs and PCB in seals from Swedish and Scottish Waters. — *Ambio* 8: 539–545.
- Borgå, K., Gabrielsen, G. W. & Skaare, J. U. 2001: Biomagnification of organochlorines along a Barents Sea food chain. — *Environmental Pollution* 113: 187–198.
- Borisov, V. I., Baranov, A. S., Valetsky, A. V. & Zakharov, V. M. 1997: Developmental stability of the mink *Mustela vison* under the impact of PCB. — *Acta Theriologica Suppl.* 4: 17–26.
- Borrell, Y. J., Pineda, H., McCarthy, I., Vasquez, E., Sanchez, J. A. & Lizana, G. B. 2004: Correlations between fitness and heterozygosity at allozyme and microsatellite loci in the Atlantic salmon, *Salmo salar* L. — *Heredity* 92: 585–593.
- Bossi, R., Rigét, F. F., Dietz, R., Sonne, C., Fauser, P., Dam, M. & Vorkamp, K. 2005: Preliminary screening of perfluorooctane sulfonate (PFOS) and other fluorochemicals in fish, birds and marine mammals from Greenland and the Faroe Islands. — *Environmental Pollution* 136: 323–329.
- Braathen, M., Derocher, A. E., Wiig, Ø., Sørmo, E. G., Lie, E., Skaare, J. U. & Jenssen, B. M. 2004: Relationships between PCBs and thyroid hormones and retinol in female and male polar bears. — *Environmental Health Perspectives* 112: 826–833.
- Clarke, G. M. 1995: Relationship between developmental stability and fitness: Application for conservation biology. — *Conservation Biology* 1: 18–24.
- Cleveland, W. S. 1979: Robust locally weighted regression and smoothing scatterplots. — *Journal of the American Statistical Association* 74(368): 829–836.
- Corsolini, S., Kannan, K., Imagawa, T., Focardi, S. & Giesy, J. P. 2002: Polychloronaphthalenes and other dioxin-like compounds in Arctic and Antarctic marine food webs. — *Environmental Science & Technology* 36: 3490–3496.
- Crespi, B. J. & Vanderkist, B. A. 1997: Fluctuating asymmetry in vestigial and functional traits of a haploid insect. — *Heredity* 79: 624–630.
- Derocher, A. E. 2005: Population ecology of polar bears at Svalbard, Norway. — *Population Ecology* 47: 267–275.
- Derocher, A. E. & Wiig, Ø. 2002: Postnatal growth in body length and mass of polar bears (*Ursus maritimus*) at Svalbard. — *Journal of Zoology* 256: 343–349.
- Derocher, A. E., Wiig, Ø. & Andersen, M. 2002: Diet composition of polar bears in Svalbard and the western Barents Sea. — *Polar Biology* 25: 448–452.
- Derocher, A. E., Wolkers, H., Colborn, T., Schlabach, M., Larsen, T. S. & Wiig, Ø. 2003: Contaminants in Svalbard polar bear samples archived since 1967 and possible population level effects. — *The Science of the Total Environment* 301: 163–174.
- Dietz, R., Heide-Jørgensen, M. P., Teilmann, J., Valentin, N. & Härkönen, T. 1991: Age determination in European Harbour seals *Phoca vitulina* L. — *Sarsia* 76: 17–21.
- Dietz, R., Rigét, F. F., Sonne-Hansen, C., Letcher, R. J., Born, E. W. & Muir, D. C. G. 2004: Seasonal and temporal trends in polychlorinated biphenyls and organochlorine pesticides in East Greenland polar bears (*Ursus maritimus*), 1990–2001. — *Science of the Total Environment* 331: 107–124.
- Dietz, R., Rigét, F. F., Sonne, C., Muir, D. C. G., Backus, S., Born, E. W., Kirkegaard, M. & Letcher, F. J. 2007: Age and seasonal variability of polybrominated diphenyl ethers in free-ranging East Greenland polar bears (*Ursus maritimus*). — *Environmental Pollution* 146: 166–173.
- Gebbink, W. A., Sonne, C., Dietz, R., Kirkegaard, M., Born, E. W., Muir, D. C. G. & Letcher, R. J. 2008: Tissue-specific composition and body burdens of organohalogen and associated metabolite contaminant in polar bears (*Ursus maritimus*) from East Greenland. — *Environmental Science and Technology*. [In press].
- Gibbs, M. & Breuker, C. 2006: Effect of larval-rearing density on adult life-history traits and developmental stability of the dorsal eyespot pattern in the speckled wood butterfly, *Pararge aegeria*. — *Entomologia Experimentalis et Applicata* 18: 41–47.
- Green, C. C. & Lochmann, S. E. 2006: Fluctuating asymmetry and condition in golden shiner (*Notemigonus crysoleucas*) and channel catfish (*Ictalurus punctatus*) reared in sublethal concentrations of isopropyl methylphosphonic acid. — *Environmental Toxicology and Chemistry* 25: 58–64.
- Grubbs, F. E. 1969: Procedures for detecting outlying observations in samples. — *Technometrics* 11: 1–21.
- Hair, J. F. Jr., Black, W. C., Babin, B. J., Anderson, R. E. & Tatham, R. L. 2006: *Multivariate data analysis*, 6th ed. — Pearson Education, Inc., New Jersey.
- Hendrickx, F., Maelfait, J.-P. & Lens, L. 2003: Relationship between fluctuating asymmetry and fitness within and between stressed and unstressed populations of the wolf spider *Pirata piraticus*. — *Journal of Evolutionary Biology* 16: 1270–1279.
- Hensel, R. J. & Sorensen, F. E. 1980: Age determination of five polar bears. — *International Conference on Bear Research and Management* 4: 93–100.
- Hoelzel, A. R., Fleischer, R. C., Campagna, C., Le Boeuf, B. J. & Alvord, G. 2002: Impact of a population bottleneck on symmetry and genetic diversity in the northern elephant seal. — *Journal of Evolutionary Biology* 15: 567–575.
- Jagoe, C. H. & Haines, T. A. 1985: Fluctuating asymmetry in fishes inhabiting acidified and unacidified lakes.

- *Canadian Journal of Zoology* 63: 130–138.
- Jones, J. C., Helliwell, P., Beekman, M., Maleszka, R. & Oldroyd, B. P. 2005: The effects of rearing temperature on developmental stability and learning and memory in the honey bee, *Apis mellifera*. — *Journal of comparative physiology A* 191: 1121–1129.
- Kirkegaard, M., Sonne, C., Leifsson, P. S., Dietz, R., Born, E. W., Letcher, R. J. & Muir, D. C. G. 2005: Histology of selected immunological organs in polar bear (*Ursus maritimus*) from East Greenland in relation to concentrations of organohalogen contaminants. — *Science of the Total Environment* 341: 119–132.
- Lajus, D. L. & Alekseev, V. R. 2004: Phenotypic variation and developmental instability of life-history traits: A theory and a case study on within-population variation of resting eggs formation in *Daphnia*. — *Journal of Limnology* 63, Suppl. 1: 37–44.
- Larsen, T. 1971: Sexual dimorphism in the molar rows of the polar bear. — *The Journal of Wildlife Management* 35: 374–377.
- Larsen, T. 1986: Population biology of the polar bear (*Ursus maritimus*) in the Svalbard area. — *Norsk Polarinstitutt Skrifter* 184: 1–53.
- Leung, B. 1998: Correcting for allometry in studies of fluctuating asymmetry and quality within samples. — *Proceedings of the Royal Society of London B* 265: 1623–1629.
- Lie, E., Larsen, H. J. S., Larsen, S., Johansen, G. M., Derocher, A. E., Lunn, N. J., Norstrom, R. J., Wiig, Ø. & Skaare, J. U. 2004: Does high organochlorine (OC) exposure impair the resistance to infection in polar bears (*Ursus maritimus*)? Part I: Effects of OCs on the humoral immunity. — *Journal of Toxicology and Environmental Health A* 67: 555–582.
- Lønø, O. 1970: The polar bear (*Ursus maritimus* Phipps) in the Svalbard area. — *Norsk Polarinstitutt Skrifter* 149: 1–115.
- Maul, J. D. & Farris, J. L. 2005: Monitoring exposure of northern cardinals, *Cardinalis cardinalis*, to cholinesterase-inhibiting pesticides: enzyme activity, reactivations, and indicators of environmental stress. — *Environmental Toxicology and Chemistry* 24: 1721–1730.
- Merilä, J. & Björklund, M. 1995: Fluctuating asymmetry and measurement error. — *Systematic Biology* 44: 97–101.
- Muir, D. C. G., Backus, S., Derocher, A. E., Dietz, R., Evans, T. J., Gabrielsen, G. W., Nagy, J., Norstrom, R. J., Sonne, C., Stirling, I., Taylor, M. K. & Letcher, R. J. 2006: Brominated flame retardants in polar bears (*Ursus maritimus*) from Alaska, the Canadian Arctic, East Greenland, and Svalbard. — *Environmental Science & Technology* 40: 449–455.
- Møller, A. P. 2006: A review of developmental instability, parasitism and disease infection, genetics and evolution. — *Infection Genetics and Evolution* 6: 133–140.
- Møller, A. P. & Swaddle, J. P. 1997: *Asymmetry, developmental stability, and evolution*. — Oxford University Press, Oxford.
- Nijhout, H. F. & Davidowitz, G. 2003: Developmental perspectives on phenotypic variation, canalization, and fluctuating asymmetry. — In: Polak, M. (ed.), *Developmental instability. Causes and consequences*: 3–13. Oxford University Press, Oxford.
- Norstrom, R. J., Belikov, S. E., Born, E. W., Garner, G. W., Malone, B., Olpinski, S., Ramsay, M. A., Schliebe, S., Stirling, I., Stishov, M. S., Taylor, M. K. & Wiig, Ø. 1998: Chlorinated hydrocarbon contaminants in polar bears from eastern Russia, North America, Greenland, and Svalbard: Biomonitoring of Arctic pollution. — *Archives of Environmental Contamination and Toxicology* 35: 354–367.
- Olsen, B. R., Reginato, A. M. & Wang, W. 2000: Bone development. — *Annual Review of Cell and Developmental Biology* 16: 191–220.
- Olsson, M. & Reutergrårdh, L. 1986: DDT and PCB pollution trends in the Swedish aquatic environment. — *Ambio* 15: 103–109.
- Palmer, A. R. 1994: Fluctuating asymmetry analyses: a primer. — In: Markow, T. A. (ed.), *Developmental instability: its origins and evolutionary implications*: 335–364. Kluwer, Dordrecht.
- Palmer, A. R. & Strobeck, C. 2003a: Fluctuating asymmetry analyses revisited. — In: Polak, M. (ed.), *Developmental instability. Causes and consequences*: 279–319. Oxford University Press, Oxford.
- Palmer, A. R. & Strobeck, C. 2003b: Fluctuating asymmetry analysis: a step-by-step example [Electronic Appendix V in: Fluctuating asymmetry analyses revisited]. — In: Polak, M. (ed.), *Developmental instability. Causes and consequences*: 279–319. Oxford University Press, Oxford.
- Pelabon, C., Hansen, T. F., Carlson, M. L. & Armbruster, W. S. 2006: Patterns of asymmetry in the twining vine *Dalechampia scandens* (Euphorbiaceae): Ontogenetic and hierarchical perspectives. — *New Phytologist* 170: 65–74.
- Pertoldi, C., Podestá, M., Loeschcke, V., Schandorff, S., Marsili, L., Mancusi, C., Nicolosi, P. & Randi, E. 2000b: Effect of the 1990 die-off in the northern Italian seas on the developmental stability of the striped dolphin *Stenella coeruleoalba* (Meyen, 1833). — *Biological Journal of the Linnean Society* 71: 61–70.
- Pertoldi, C., Bach, L. A., Barker, J. S. F., Lundberg, P. & Loeschcke, V. 2007: The consequences of the variance-mean rescaling effect on effective population size. — *Oikos* 116: 769–774.
- Pertoldi, C., Bach, L. A., Madsen, A. B., Randi, E. & Loeschcke, V. 2003: Morphological variability and developmental instability in subpopulations of the Eurasian badger (*Meles meles*) in Denmark. — *Journal of Biogeography* 30: 949–958.
- Pertoldi, C., Loeschcke, V., Braun, A., Madsen, A. B. & Randi, E. 2000a: Craniometrical variability and developmental stability. Two useful tools for assessing the population viability of Eurasian otter (*Lutra lutra*) populations in Europe. — *Biological Journal of the Linnean Society* 70: 309–323.
- Pertoldi, C., Loeschcke, V., Madsen, A. B. & Randi, E. 1997: Developmental stability in the Eurasian Otter (*Lutra lutra*) in Denmark. — *Annales Zoologici Fennici* 34: 187–196.

- Petavy, G., David, J. R., Debat, V., Pertoldi, C. & Moreteau, B. 2006: Phenotypic and genetic variability of sternopleural bristle number in *Drosophila melanogaster* under thermal stress: developmental instability and anti-symmetry. — *Evolutionary Ecology Research* 8(1): 149–167.
- Polischuk, S. C., Norstrom, R. J. & Ramsay, M. A. 2002: Body burdens and tissue concentrations of organochlorines in polar bears (*Ursus maritimus*) vary during seasonal fasts. — *Environmental Pollution* 118: 29–39.
- Pravosudov, V. V. & Kitaysky, A. S. 2006: Effects of nutritional restrictions during post-hatching development on adrenocortical function in western scrub-jays (*Aphelocoma californica*). — *General and Comparative Endocrinology* 145(1): 25–31.
- Prince, M. E. & Silva, M. 2002: On the determinants of census area: implications for mammalian macroecological patterns. — *Global Ecology & Biogeography* 11: 401–409.
- Ramsay, M. A. & Stirling, I. 1988: Reproductive biology and ecology of female polar bears (*Ursus maritimus*). — *Journal of Zoology* 214: 601–634.
- Regehr, E. V., Lunn, N. J., Amstrup, S. C. & Stirling, I. 2005: Population decline of polar bears in Western Hudson Bay in relation to climatic warming. — In: *16th Biennial Conference on Marine Mammals, San Diego, California, December 12th–16th, 2005*: 233–234. Society for Marine Mammalogy.
- Riget, F. 2005: Tidsmæssig udvikling af kontaminanter i grønlandske dyr — en statusrapport. *Arbejdsrapport fra DMU (Danmarks Miljøundersøgelse)*, Nr. 228.
- Rosing-Asvid, A., Born, E. W. & Kingsley, M. C. S. 2002: Age at sexual maturity of males and timing of the mating season of polar bears (*Ursus maritimus*) in Greenland. — *Polar Biology* 25(12): 878–883.
- Sandala, G. M., Sonne-Hansen, C., Dietz, R., Muir, D. C. G., Valters, K., Bennett, E. R., Born, E. W. & Letcher, R. J. 2004: Hydroxylated and methyl sulfone PCB metabolites in adipose and whole blood of polar bear (*Ursus maritimus*) from East Greenland. — *Science of the Total Environment* 331: 125–141.
- Schandorff, S. 1997: Developmental stability and skull lesions in the harbour seal (*Phoca vitulina*) in the 19th and 20th centuries. — *Annales Zoologici Fennici* 34: 151–166.
- Siegel, M. I. & Mooney, M. P. 1987: Perinatal stress and increased fluctuating asymmetry of dental calcium in the laboratory rat. — *American Journal of Physical Anthropology* 73: 267–270.
- Siegel, M. I., Doyle, W. J. & Kelley, C. 1977a: Heat stress, fluctuating asymmetry and prenatal selection in the laboratory rat. — *American Journal of Physical Anthropology* 46: 121–126.
- Siegel, P., Siegel, M. I., Krimmer, E. C., Doyle, W. J. & Barry, H. III 1977b: Fluctuating dental asymmetry as an indicator of the stressful prenatal effects of Δ^9 -tetrahydrocannabinol in the laboratory rat. — *Toxicology and applied pharmacology* 42: 339–344.
- Skaare, J. U., Bernhoft, A., Derocher, A. E., Gabrielsen, G. W., Goksøyr, A., Henriksen, E., Larsen, H. J., Lie, E. & Wiig, Ø. 2000: Organochlorines in top predators at Svalbard — occurrence, levels and effects. — *Toxicology Letters* 112–113: 103–109.
- Smithwick, M., Norstrom, R. J., Mabury, S. A., Solomon, K., Evans, T. J., Stirling, I., Taylor, M. K. & Muir, D. C. G. 2006: Temporal trends of perfluoroalkyl contaminants in polar bears (*Ursus maritimus*) from two locations in the North American Arctic, 1972–2002. — *Environmental Science & Technology* 40: 1139–1143.
- Sonne, C., Dietz, R., Leifsson, P. S., Born, E. W., Kirkegaard, M., Riget, F. F., Letcher, R., Muir, D. C. G. & Hyldstrup, L. 2005b: Do organochlorine contaminants contribute to histopathology in liver from East Greenland polar bears (*Ursus maritimus*)? — *Environmental Health Perspectives* 113: 1569–1574.
- Sonne, C., Dietz, R., Leifsson, P. S., Born, E. W., Kirkegaard, M., Letcher, R. J., Muir, D. C. G., Riget, F. F. & Hyldstrup, L. 2006a: Are organohalogen contaminants a co-factor in the development of renal lesions in East Greenland polar bears (*Ursus maritimus*)? — *Environmental Toxicology and Chemistry* 25: 1551–1557.
- Sonne, C., Leifsson, P. S., Dietz, R., Born, E. W., Letcher, R. J., Hyldstrup, L., Riget, F. F., Kirkegaard, M. & Muir, D. C. G. 2006b: Xenoendocrine pollutants may reduce size of sexual organs in East Greenland polar bears (*Ursus maritimus*). — *Environmental Science & Technology* 40: 5668–5674.
- Sonne, C., Dietz, R., Born, E. W., Riget, F. F., Kirkegaard, M., Hyldstrup, L., Letcher, R. J. & Muir, D. C. G. 2004: Is bone mineral composition disrupted by organochlorines in East Greenland polar bears (*Ursus maritimus*)? — *Environmental Health Perspectives* 112: 1711–1716.
- Sonne, C., Riget, F. F., Dietz, R., Kirkegaard, M., Born, E. W., Letcher, R. & Muir, D. C. G. 2005a: Trends in fluctuating asymmetry in East Greenland polar bears (*Ursus maritimus*) from 1892 to 2002 in relation to organohalogen pollution. — *Science of the Total Environment* 341: 81–96.
- Sonne, C., Rigét, F. F., Dietz, R., Wiig, Ø., Kirkegaard, M. & Born, E. W. 2007: Skull pathology in East Greenland and Svalbard polar bears (*Ursus maritimus*) during 1892 to 2002 in relation to organochlorine pollution. — *Science of the Total Environment* 372: 554–561.
- Stige, L. C. 2004: *Fluctuating asymmetry: a marker of individual genetic or phenotypic quality?* — Ph.D. thesis, Department of Biology, University of Oslo.
- Valetsky, A. V., Dmitrieva, I. L., Krushinskaya, N. L. & Zakharov, V. M. 1997: Social stress impact on developmental stability of laboratory rat *Rattus norvegicus*. — *Acta Theriologica* Suppl. 4: 27–32.
- Verreault, J., Muir, D. C. G., Norstrom, R. J., Stirling, I., Fisk, A. T., Gabrielsen, G. W., Derocher, A. E., Evans, T. J., Dietz, R., Sonne, C., Sandala, G. M., Gebbink, W., Riget, F. F., Born, E. W., Taylor, M. K., Nagy, J. & Letcher, R. J. 2005: Chlorinated hydrocarbon contaminants and metabolites in polar bear (*Ursus maritimus*) from Alaska, Canada, East Greenland, and Svalbard: 1996–2002. — *Science of the Total Environment* 351–352: 369–390.
- Vieth, R., Ladak, Y. & Walfish, P. G. 2003: Age-related

- changes in the 25-hydroxyvitamin D versus parathyroid hormone relationship suggest a different reason why older adults require more vitamin D. — *The Journal of Clinical Endocrinology & Metabolism* 88: 185–191.
- Vilicsics, F., Solymos, P. & Hornung, E. 2005: Measuring fluctuating asymmetry of the terrestrial isopod *Trachelipus rathkii* (Crustacea: Isopoda, Oniscidea). — *European Journal of Soil Biology* 41: 85–90.
- Voigt, C. C., Heckel, G. & Mayer, F. 2005: Sexual selection favours small and symmetric males in the polygynous greater sac-winged bat *Saccopteryx bilineata* (Emballonuridae, Chiroptera). — *Behavioral Ecology and Sociobiology* 57: 457–464.
- Wiig, Ø. 2005: Are polar bears threatened? — *Science* 309: 1814–1815.
- Wiig, Ø., Derocher, A. E., Cronin, M. M. & Skaare, J. U. 1998: Female pseudohermaphrodite polar bears at Svalbard. — *Journal of Wildlife Diseases* 34: 792–796.
- Zakharov, V. M. 1992: Population phenogenetics: Analysis of developmental stability in natural populations. — *Acta Zoologica Fennica* 191: 7–30.
- Zakharov, V. M. & Yablokov, A. V. 1990: Skull asymmetry in the Baltic grey seal: effects of environmental pollution. — *Ambio* 19: 266–269.
- Zakharov, V. M., Valetsky, A. V. & Yablokov, A. V. 1997a: Dynamics of developmental stability of seals and pollution in the Baltic Sea. — *Acta Theriologica* Suppl. 4: 9–16.
- Zakharov, V. M., Demin, D. V., Baranov, A. S., Borisov, V. I., Valetsky, A. V. & Sheftel, B. I. 1997b: Developmental stability and population dynamics of shrews *Sorex* in central Siberia. — *Acta Theriologica* Suppl. 4: 41–48.